

POLYVINYL ALCOHOL-ALGINATE FERRO PHOTO GELS FOR  
MERCURY(II) REMOVAL

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*This entire research is dedicated with deepest love to my dearest parents, brother  
and sisters.*

*Thank you for the never ending love, trust, understanding, support, and motivation.*

*May Allah give you the best reward and may the blessings of Allah be upon you.*

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## ABSTRACT

Mercury (Hg(II)) is considered as one of the most noxious heavy metals due to its high toxicity and probability of bioaccumulation in human body. Adsorption is one of the most commonly used technique to treat Hg(II) in wastewater but it requires a two stage process where the toxic Hg(II) is not being converted into non-toxic medium but need to be recovered. Thus photocatalytic process was introduced where the photocatalyst reacts by converting the toxins into toxic-free wastes; henceforth there is no necessity for additional disposal. In this study nano-sized maghemite ( $\gamma\text{-Fe}_2\text{O}_3$ ) embedded in polyvinyl alcohol (PVA) and alginate matrix was used as photocatalyst to remove Hg(II). Besides being a photocatalyst, the  $\gamma\text{-Fe}_2\text{O}_3$  nanoparticles possess supermagnetism properties which enabled the beads to be easily recovered with the application of an external magnetic field. The influences of pH, initial concentration and photocatalyst dosage on Hg(II) removal were also investigated. The photocatalyst beads were then used for the reduction of Hg(II) in aqueous solution both under sunlight and away from sunlight. The synthesized maghemite nanoparticles were characterized using transmission electron microscopy, x-ray diffraction and vibrating sample magnetometer; and the size distribution of the beads were determined. The current results revealed that 96% of Hg(II) was reduced in four hours under sunlight. However, when the experiment was done in the dark, the percentage of Hg(II) reduction achieved was only 5%. The low reduction rate was due to the minimal absorption activity of Hg(II) onto the beads surface. In addition, the maximum Hg(II) reduction was found at pH 11 whilst the photocatalyst dosage was kept at 16% (v/v). An 8% (v/v) dosage of maghemite nanoparticles loading was found to be capable only to reduce until 67% of Hg(II), while Hg(II) reduction performance was not significantly improved when 24% (v/v) of photocatalyst dosage was used. Excessive addition of catalyst dosage increased the active sites on the beads surface but it also blocked some sunlight illumination as the voluminous load of photocatalyst clogged the reaction region thus reducing the photon availability to be absorbed. At a fixed optimum parameters, it was revealed that increasing the initial concentration of Hg(II) degraded the reduction capability because the photons path length into the solution reduced as concentration of Hg(II) increased. Field emission scanning electron microscopy images and energy dispersive x-ray showed that the beads possessed significant porosity structure that greatly supported mass movement of Hg(II) inside the beads. The maghemite embedded PVA-alginate beads towards reduction of Hg(II) strongly fitted Langmuir-Hinshelwood kinetics model with correlation coefficient,  $R^2$  value of 0.9771. In conclusion, this study proved that the  $\gamma\text{-Fe}_2\text{O}_3$ -PVA-alginate beads are applicable in reducing and treating Hg(II) in water.

## ABSTRAK

Raksa (Hg(II)) dianggap salah satu daripada logam berat paling berbahaya disebabkan sangat bertoksik dan kebarangkalian bioakumulasi di dalam badan manusia. Penjerapan adalah salah satu teknik yang selalu digunakan untuk merawat Hg(II) dalam sisa air tetapi ia memerlukan proses dua peringkat di mana Hg(II) bertoksik tidak ditukar kepada bahan tidak bertoksik tetapi perlu untuk diperolehi semula. Oleh itu proses fotopemangkin telah diperkenalkan di mana fotomangkin bertindak balas dengan menukarkan toksin kepada sisa bebas daripada toksik; sekaligus tiada keperluan untuk pelupusan tambahan. Di dalam kajian ini, maghemit bersaiz nano ( $\gamma\text{-Fe}_2\text{O}_3$ ) yang telah dimasukkan ke dalam polivinil alkohol (PVA) dan matriks alginat telah digunakan sebagai fotomangkin untuk menyingkirkan Hg(II). Selain bertindak sebagai fotomangkin, partikel bersaiz nano  $\gamma\text{-Fe}_2\text{O}_3$  ini memiliki sifat superkemagnetan yang membolehkan manik itu mudah diperolehi semula dengan menggunakan medan magnet luaran. Pengaruh daripada pH, kepekatan awal dan dos fotomangkin terhadap penyingkiran Hg(II) juga telah disiasat. Manik fotomangkin ini kemudiannya digunakan untuk merawat Hg(II) di dalam larutan air di bawah sinaran matahari dan tanpa cahaya matahari. Partikel bersaiz nano maghemit yang diteliti disediakan telah dicirikan dengan menggunakan mikroskopi pancaran elektron, pembelauan sinar-x dan magnetometer sampel getaran; dan taburan saiz manik telah ditentukan. Keputusan menunjukkan bahawa 96% Hg(II) telah dirawat dalam empat jam di bawah sinaran matahari. Walau bagaimanapun, apabila eksperimen dilakukan di dalam gelap, peratusan penurunan Hg(II) hanya mencecah 5%. Kadar penurunan yang rendah adalah disebabkan oleh aktiviti penyerapan Hg(II) yang minima ke atas permukaan manik-manik. Selain itu, didapati bahawa pengurangan maksima Hg(II) adalah pada pH 11 di mana dos fotomangkin digunakan pada 16% (v/v). Dos maghemit zarah nano pada 8% (v/v) hanya mampu mengurangkan sehingga 67% Hg(II) sementara prestasi penurunan Hg(II) tidak banyak berubah apabila 24% (v/v) dos fotomangkin digunakan. Penggunaan berlebihan dos pemangkin akan menambah ruang aktif di permukaan manik, tetapi ia juga akan menghalang kemasukan cahaya matahari kerana fotomangkin yang berlebihan menyumbat kawasan tindakbalas sekaligus mengurangkan penyerapan foton. Pada parameter optimum yang telah ditetapkan, didapati dengan meningkatkan kepekatan awal Hg(II), keupayaan penurunan merosot kerana jarak laluan foton di dalam larutan berkurangan setelah kepekatan Hg(II) ditingkatkan. Imej mikroskopi pengimbas pancaran medan elektron dan tenaga serakan sinar-x menunjukkan bahawa manik memiliki struktur berliang ketara yang membantu pergerakan jisim Hg(II) di dalam manik. Manik-manik PVA-alginat berisi maghemit menuju kepada penurunan Hg(II) melengkap model kinetik Langmuir-Hinshelwood dengan pekali korelasi,  $R^2$  bernilai 0.9771. Kesimpulannya, kajian ini membuktikan bahawa manik  $\gamma\text{-Fe}_2\text{O}_3$ -PVA-alginat boleh digunakan dalam mengurangkan dan merawat Hg(II) di dalam air.

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## LIST OF ABBREVIATIONS

Aliquat 336	-	Tri-Octyl Methylammonium Chloride
Au-Fe <sub>3</sub> O <sub>4</sub>	-	Gold-Ferrous-Ferric Oxide
Ag	-	Argentum / Silver
BDETH <sub>2</sub>	-	Benzene-1, 3-Diamidoethanethiol
CdS	-	Cadmium Sulfide
CH <sub>3</sub> Hg	-	Monomethylmercury
(CH <sub>3</sub> ) <sub>2</sub> Hg	-	Dimethylmercury
Co-Mn	-	Cobalt-Manganese
Cr(III)	-	Trivalent Chromium
Cr(VI)	-	Hexavalent Chromium
CrO <sub>4</sub> <sup>-</sup>	-	Chromate ion
CuCrO <sub>2</sub>	-	Copper-Chromium Oxide
CVAAS	-	Cold Vapor Atomic Absorption Spectrometer
DBT	-	Dibenzothiophene
DNA	-	Deoxyribonucleic Acid
DOE	-	Department of Environment
EDTA	-	Ethylenediamine Tetraacetate
EDX	-	Energy Dispersive X-Ray
FESEM	-	Field Emission Scanning Microscopy
Fe <sub>2</sub> O <sub>3</sub>	-	Maghemite
Fe <sub>3</sub> O <sub>4</sub>	-	Magnetite / Iron(II) Di-Iron(III) Oxide
Fe(OH) <sub>3</sub>	-	Iron(III) Oxide-Hydroxide
FeTiO <sub>3</sub>	-	Iron Titanium Oxide
H <sub>2</sub>	-	Hydrogen
Hg <sup>0</sup>	-	Elemental Mercury
Hg <sup>2+</sup>	-	Mercuric ion
Hg(II)	-	Divalent Mercury

HgCl <sub>2</sub>	-	Mercury(II) Chloride
Hg(OH) <sub>2</sub>	-	Mercury(II) Hydroxide
Hg(NO <sub>3</sub> ) <sub>2</sub>	-	Mercury(II) Nitrate
HVAC	-	Heating, Ventilation, and Air Conditioning
L-H	-	Langmuir-Hinshelwood
M-M	-	Mannuronate-Mannuronate Blocks
MgO	-	Magnesium Oxide
MN	-	Maghemite Nanoparticles
MRI	-	Magnetic Resonance Imaging
NiPc	-	Nickel Phthalocyanine
PAD	-	Photo-Assisted Deposition
PEI	-	Polyethyleneimine
POM	-	Polyoxometalates
PW <sub>12</sub> O <sub>40</sub> <sup>3-</sup>	-	1-Equiv reduced tungstophosphate
ppb	-	Parts per billion
ppm	-	Parts per million
PPEI	-	Phosphonomethylated Polyethyleneimine
PVA	-	Polyvinyl Alcohol
PyDETH <sub>2</sub>	-	Pyridine-2,6-Diamineethanethiol
ROS	-	Reactive Oxygen Species
SDTC	-	Sodium Dimethyldithiocarbamate
SH-Fe <sub>3</sub> O <sub>4</sub> -NMPs	-	Sulfur-containing ligands mercapto-functionalized Nano-Fe <sub>3</sub> O <sub>4</sub> Magnetic Polymers
SIRIM	-	Standards and Industrial Research Institute of Malaysia
SiW <sub>12</sub> O <sub>40</sub> <sup>4-</sup>	-	One-electron-reduced 12-tungstosilicate
STC	-	Sodium Thiocarbonate
TEM	-	Transmission Electron Microscopy
TiO <sub>2</sub>	-	Titanium Dioxide
TMT	-	Trimercapto-1, 3, 5-Triazine
UV	-	Ultraviolet
UV-vis	-	Ultraviolet-visible
VOC	-	Volatile Organic Compounds
VSM	-	Vibrating Sample Magnetometer

V-TiO <sub>2</sub>	-	Vanadium-doped Titanium Dioxide
WO <sub>3</sub>	-	Tungsten Oxide
WS <sub>2</sub>	-	Tungsten Sulfide
XRD	-	X-Ray Diffractometer
ZnO	-	Zinc Oxide
ZnS	-	Zinc Sulfide
4-BPDB	-	1,4-Bis(4-Pyridyl)-2,3-Diaza-1,3-Butadiene
γ-Fe <sub>2</sub> O <sub>3</sub>	-	Maghemite
α-Fe <sub>2</sub> O <sub>3</sub>	-	Hematite / Iron(III) Oxide

## LIST OF SYMBOLS

$C_0$	-	Initial concentration of solution (mg/L)
$C$	-	Concentration of solution (mg/L)
$C_w$	-	Concentration of water molecules
$d$	-	Diameter
$E_{CB}$	-	Conduction band-edges
$E_{VB}$	-	Valence band-edges
$eV$	-	Electronvolt
$e^-$	-	Electron
$h$	-	Initial sorption rate (mg/L h)
$h\nu$	-	Surplus energy, E
$g$	-	Gram
$k$	-	Rate constant of pseudo second-order equation (L/mg h)
$k_r$	-	Reaction rate constant (mg/L min)
$k_{app}$	-	Reaction rate ( $\text{min}^{-1}$ )
$K_{LH}$	-	Adsorption constant of the reactant (L/mg)
$K_w$	-	Solvent adsorption constant
$L$	-	Litre
$m^2$	-	Square metre
$mg$	-	Miligram
$min$	-	Minute
$ml$	-	Mililitre
$mm$	-	Milimetre
$ng$	-	Nanogram
$nm$	-	Nanometre
$r$	-	Reduction rate (mg/L min)
$R^2$	-	Correlation coefficient
$r_0$	-	Initial rate of photocatalytic reduction



t	-	Illumination time (minute)
w/v	-	Weight/volume
q <sub>e</sub>	-	Solute loading on the solid phase (mg/g)
λ	-	Wavelength
Å	-	Ångström or 0.1 nm
μg	-	Microgram

## CHAPTER 1

### INTRODUCTION

#### 1.1 Overview

On 1956, an irresponsible release of methylmercury by a chemical company into the sea caused a deathly syndrome in Japan. This severe mercury poisoning which is called Minamata disease had haunted the Japanese people for 36 years as acute effects continued to rise. Some of them experienced hearing and speech damage, hands and feet numbness, insanity, paralysis, effect on the foetus in the womb and even death (Zhang *et al.*, 2004).

In Malaysia, mercury wastes are mainly originate from agricultural pesticides and they also emerge from the chlorine-alkali industry. Moreover, mercury is used as a catalyst in the chemical and petrochemical industries, used in electrical apparatus, cosmetics, thermometers, gauges, batteries, painting and coating industries, mining, extractive metallurgy and many other industries (Parham *et al.*, 2012). This environmental pollution stimulates concerns about the dangers posed to human being, thus numerous efforts have been taken to handle this problem, essentially aiming at cost effectiveness. Hence, Malaysian Department of Environment (DOE)

sets tolerance limit release for Hg(II) at 0.005 mg/L for standard A and 0.05 mg/L for standard B.

Mercury is considered as one of the most noxious heavy metals acknowledged by mankind, due to high toxicity and probability of bioaccumulation in body. It forms different salts with anionic mechanisms in water and can change from one form to another in various diverse aquatic environments (Parham *et al.*, 2012). Some severe and chronic signs caused by mercury poisoning are metallic taste, inflammation in mouth, kidney collapse and excessive salivation (Parham *et al.*, 2012).

Many methods have been used to remove Hg(II) from wastewater such as ion exchange, precipitation as sulphide, membrane filtration, ion exchange, electrodeposition, coagulation, reverse osmosis, electro-deposition, ultrafiltration and adsorption (Li *et al.*, 2008). One of the techniques that is most widely used to date is adsorption. Adsorption method is chosen because of its high efficiency, availability of different adsorbents, sorbent materials can be generated and recycled, easy handling and most importantly cost effectiveness (Xiong *et al.*, 2009). Unfortunately, a secondary method need to be introduced to desorb the heavy metals before disposing the adsorbents.

In a recent study, maghemite nanoparticles ( $\gamma\text{-Fe}_2\text{O}_3$ ) were proven as an excellent adsorbent for removing heavy metals due to the vast surface area which improves the adsorption, many methods to synthesize maghemite and at the same time, it is commercially available. Also, easy separation of metal and magnetic adsorbent from treated water, and ultimately, no secondary waste will be produced (Tuutijärvi *et al.*, 2009).

By applying external magnetism to the magnetic particles, excellent disintegration will be accomplished (Hu *et al.*, 2005). Accordingly, studies have been initiated by combining magnetic separation with biosorption technique. This method is well known of treating many heavy metals in wastewater in a short time, while no contaminants will be formed (Rocher *et al.*, 2008). In addition, several studies have recorded positive results; encapsulation of maghemite nanoparticles ( $\gamma\text{-Fe}_2\text{O}_3$ ) in sodium alginate were used to remove Pb(II) and to reduce Cr(VI) to less toxic Cr(III),  $\gamma\text{-Fe}_2\text{O}_3$  was incorporated into PVA and alginate matrix photocatalyst to reduce Cr(VI) to Cr(III) and was used to removal of  $\text{Cs}^+$  from radioactive waste water (Idris *et al.*, 2010, 2012; Majidnia and Idris, 2015). Likewise, Cu(II) was also successfully removed by maghemite nanoparticles doped with cobalt entrapped in PVA-alginate beads (Wong *et al.*, 2014).

The maghemite nanoparticles ( $\gamma\text{-Fe}_2\text{O}_3$ ) which are embedded in binding materials such as alginate which has been used to encapsulate magnetic particles to form into beads, can be recycled (Ngomsik *et al.*, 2006, 2009). Being hydrophilic, biodegradable, inexpensive, non-toxic, naturally available and with existence of binding sites contribute from its carboxylate functions, have favour alginate over other materials. On the other hand, Liu *et al.* (2008) has found that drug delivery ferrogels with magnetic-stimuli performance is very much influenced on the concentration of PVA. PVA is nontoxic, inexpensive, robust and possesses an uncomplicated chemical structure which allows alteration to be made by executing a chemical reaction, and it is combined with alginate to reduce PVA's agglomeration.

Besides that, there are various semiconductor materials that are chosen in photocatalytic practice, for example titanium dioxide ( $\text{TiO}_2$ ), zinc oxide (ZnO),  $\text{WO}_3$  (tungsten oxide) and polyoxometalate anions (POM). Based on Ullah and Dutta (2008) findings, these semiconductors have individual band gap in ultraviolet (UV) region, thus aiding photocatalysis with radiance from UV emission.

## 1.2 Problem Statement

Many techniques were developed to treat wastewater; among them were absorption and photocatalyst reduction. Parham *et al.* (2012) used altered magnetic iron oxide nanoparticles with 2-mercaptobenzothiazole as an adsorbent for Hg(II) removal and results revealed a complete removal of mercury ions within 4 minutes, and the modified magnetic iron oxide nanoparticles can be reutilized three times without decreasing adsorption efficiency. However, in adsorption, the impurities are being absorbed into the bioadsorbents but then again are not being converted into non-toxic medium. Thus photocatalytic process was introduced where the photocatalyst reacts by converting the toxins into toxic-free wastes; henceforth there is no necessity for additional disposal. Recently, magnetic nanoparticles iron oxide ( $\gamma\text{-Fe}_2\text{O}_3$ ) was used as a photocatalyst and it has an excellent efficiency in converting hazardous metal ion such as chromium(VI) to chromium(III) (Idris *et al.*, 2010, 2011). The photocatalyst reaction comes from electrons couples which photoproduction from the valence and conduction ranges, with absorption of UV radiation with energy equivalent or greater than the range. However the use of  $\gamma\text{-Fe}_2\text{O}_3$  was never applied for the removal of Hg(II).

Thus in this study, maghemite nanoparticles ( $\gamma\text{-Fe}_2\text{O}_3$ ) which is the photocatalyst is embedded in the alginate-PVA matrix is used to treat wastewater containing Hg(II). Photocatalytic experiments will be performed in the dark and under sunlight and several some parameters such as pH, photocatalytic dosage, initial feed concentration were investigated.

### 1.3 Objectives

The main objective of this study is to remove Hg(II) from aqueous solution using maghemite  $\gamma\text{-Fe}_2\text{O}_3$  embedded in PVA-alginate. In order to focus on this objective, the following need to be addressed:

- i) To prepare the  $\gamma\text{-Fe}_2\text{O}_3$  PVA-alginate beads
- ii) To study its physicochemical properties
- iii) To test the performance of the catalyst on removal of Hg(II)
- iv) To study the kinetics of the photocatalytic reaction

### 1.4 Scope of The Study

In order to achieve the objectives mentioned, the encompassed work includes the following:

- i) The maghemite nanoparticles were synthesized using the coprecipitation method and then embedded in the PVA-alginate matrix.
- ii) The synthesized maghemite nanoparticles were characterized using TEM (Transmission electron microscopy), XRD (X-ray diffraction) and VSM (Vibrating Sample Magnetometer); and the size distribution of the beads were determined.
- iii) The photocatalytic experiments were then performed in darkness and sunlight, using maghemite nanoparticles and without using maghemite nanoparticles.
- iv) The photocatalytic experiments were performed at various pH ranging from 2 until 13.

- v) The influence of photocatalyst dosage on Hg(II) removal was determined; where maghemite  $\gamma\text{-Fe}_2\text{O}_3$  dosages were varied at 8%, 16% and 24%.
- vi) Hg(II) initial concentration was determined and varied at 25 ppm, 50 ppm, 75 ppm and 100 ppm. The effect of Hg(II) initial concentration on the Hg(II) removal was analysed.
- vii) Hg(II) elements in the beads before and after the experiment was analysed with CVAAS (Cold-vapour atomic absorption spectrometry).
- viii) Kinetic analysis of photoreduction was determined for the optimum photocatalytic.

### 1.5 Significance of The Study

Earlier studies (Idris *et al.*, 2010, 2012) have used ferrophoto gel beads with embedded maghemite nanoparticles in the photocatalyst process to treat Cr(VI). López-Muñoz *et al.* (2011) used  $\text{TiO}_2$  in his photocatalyst study, while other techniques on Hg(II) removal on the other hand did not apply photocatalytic process. Parham *et al.* (2012) removed mercury from wastewater with magnetic iron oxide nanoparticles modified with 2-mercaptobenzothiazole using adsorption process. Thus in this study, effectiveness of  $\gamma\text{-Fe}_2\text{O}_3$  embedded PVA-alginate beads in removal of Hg(II) using photocatalytic process is investigated for the first time.

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