

Multimedia Environmental Fate and Transport Model of dichlorodiphenyltrichloroethane (DDT): Case Study Sayong River Watershed .

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Abstract

This study describes the development of a multimedia environmental fate and transport model of dichlorodiphenyltrichloroethane (DDT) at Sayong River Watershed. Based on the latest estimated DDT emission, the DDT concentrations in air, soil, water, and sediment as well as the transfer flux were simulated under the equilibrium and steady-state assumption. Model predictions suggest that soil and sediment was the dominant sink of DDT. The results showed that the model predicted was generally good agreement with field data. Compared with degradation reaction, advection outflow was more important in removal processes of DDT. Sensitivities of the model estimates to input parameters were tested. The result showed that vapour pressure (P_s), chemical solubility (C_s), and absolute temperature (T) of DDT were the most influential parameters for the model output. The model output—concentrations of DDT in multimedia environment is very important as it can be used in future for human exposure and risk assessment of organochlorine pesticides (OCPs) at Sayong River Basin.

Keywords

Multimedia environmental fate and transport model; dichlorodiphenyltrichloroethane (DDT); modeled concentration; transfer fluxes.

1. INTRODUCTION

Over the past, environmental deterioration and natural resource destruction occurred caused by agricultural and industrial development (Sultana *et al.*, 2014). In chemical management and environmental decision-making, it is very important to assess the regional ecological and human health risk of chemicals released into the environment (Liu *et al.*, 2014). For this purpose sampling analysis methods are always used by researcher, but unfortunately it cannot practically capture the dynamic behavior of contaminants because these methods are costly, time-consuming and laborious and it is not practical for real-time prediction emergent events that might pose risks (Wang, Feng, Gao, Ren, & Li, 2012). Thus, a new tool called multimedia environmental modeling (MEM) has been introduced to predict the level distribution of a contaminant in all connected environmental (Luo, Gao, & Yang, 2007). Therefore, in Decision Support System (DSS) for chemical risk assessment, this type model can be applied which becoming urgent need for management of Persistent Organic Pollutants (POPs) in Malaysia. The most important thing, this study is providing a useful tool for chemical risk assessment especially on a watershed scale in Malaysia in future.

As one of agricultural country, pesticides are widely used in Malaysia. In prevention of harmful effects caused by pests, pesticide which is constitute a diverse class of chemicals extensively used and its accumulated in soil (Ene Antoaneta 2012). Among the large numbers of pesticides, Organochlorine Pesticides (OCPs) included as an ubiquitous Persistent Organic Pollutants (POPs) and have been a major environmental issue, drawing extensive attention from environmental scientist and public (Kim, Lee, Lee, Jacobs Jr, & Lee, 2015). Some Pesticides (OCPs) such as dichlorodiphenyltrichloroethane (DDT), dieldrin and hexachlorocyclohexanes (HCH) have been withdrawn or banned in many countries for environmental reasons and public health (Mahugija, Henkelmann, & Schramm, 2014). Because of good effect in controlling insects and low cost, a number of Organochlorine Pesticides (OCPs) are still in use in South Asian Countries even though prohibitions on the uses of them have been implemented in developing nations (Usman *et*

al., 2014). Because the ability of Organochlorine Pesticides (OCPs) can accumulated mainly in animal tissue and enter the food chain, thus Organochlorine Pesticides (OCPs) are considered toxic substance (Luzardo *et al.*, 2012). Their occurrence in remote areas is attributed to transport of chemicals from places where they are still in use because Organochlorine Pesticides (OCPs) are characterized by long range atmospheric transport potential (Usman *et al.*, 2014). Thus, Organochlorine Pesticides (OCPs) were found widespread in the environmental media, such as soil, water, suspended particulate matter (SPM), sediment, atmosphere and organisms (Yu *et al.*, 2014).

Among many types of Multimedia Environmental Models (MEM), multimedia fugacity model has been widely used to describe the environmental behaviour of organic pollutants in local environment, regional and global environments (Xiangzhen *et al.*, 2014). The concept of fugacity and mass balance principle are used to describe the partitioning processes in different environmental compartments and predict the concentration level, distribution and persistence of the chemicals (Liu *et al.*, 2014).

Sayong River is one of the most important river in Malaysia because it was a water sources for South Malaysia and Singapore. The objective of this study was to develop a multimedia environmental fate and transport model to estimate the environmental distribution and transfer fluxes of dichlorodiphenyltrichloroethane (DDT) in air, soil, water and sediment at Sayong River watershed. Monitoring study was conducted on December 2014 and February 2015. The measured distributions of dichlorodiphenyltrichloroethane (DDT) were compared with the modeled concentration results using log difference for model validation. The model was also assessed by using sensitivity analysis to determine the most influential parameters for the model. The results from this study were expected to be useful for local government agencies to manage dichlorodiphenyltrichloroethane (DDT) at the watershed.

2. METHODOLOGY

2.2 Study Area and Segmentation

Sayong River is one of the 12 tributaries in the Johor River Basin. The Sayong River has a watershed area of about 480 238 km². There are 12 tributaries in the Sayong River watershed. In order to distinguish different types of study areas, Sayong River watershed was divided into 3 sections: upstream, midstream, downstream using Geographical Information System (GIS) analysis. In general, the economic activities in the 3 sections are same, with mainly having palm oil plantation activities. Sayong River Basin is important to be managed effectively because it is the source of water intake in Johor. Monitoring activities were carried out on December 2014 and January 2015 to provide the DDT data in air, soil, water and sediment at Sayong River watershed. Study area and sampling stations were presented in Figure 2-1. Those data was used in model validation.

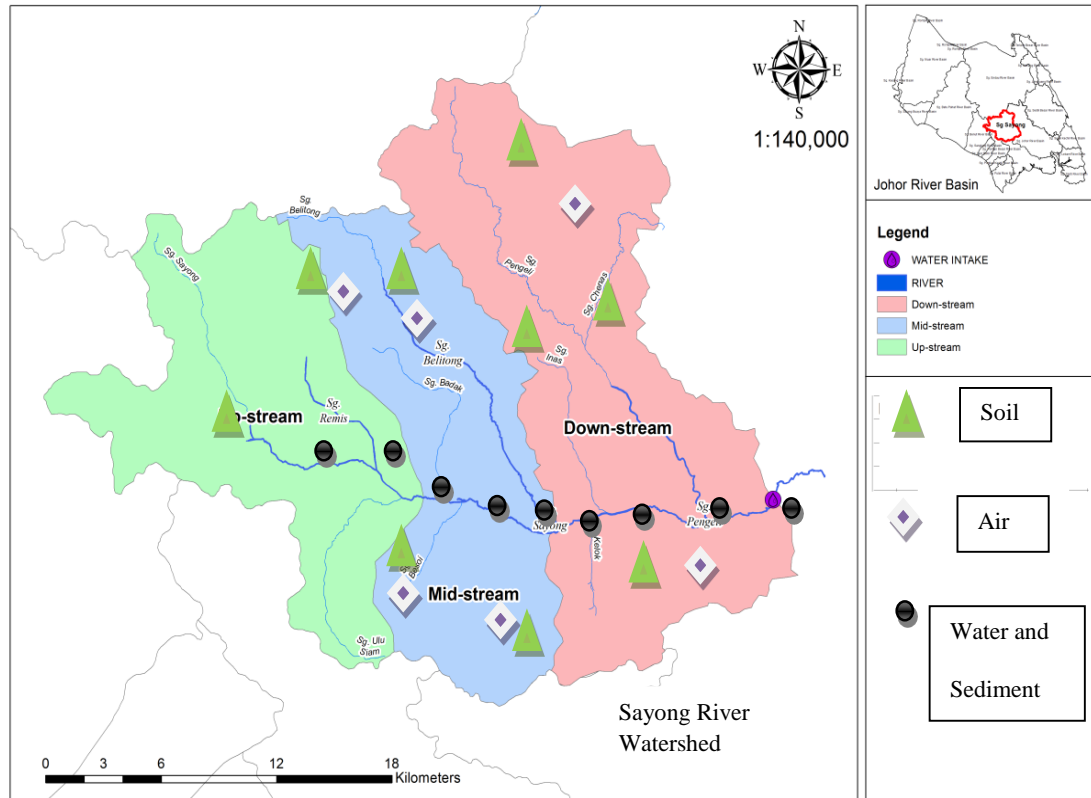


Figure 2-1: Sampling stations

2.3 Model Framework

Based on approach of Mackay, 2001, a fugacity model was developed to describe the distribution, degradation reaction and advection loss processes of DDT at Sayong River Basin as well as its different sections.

The processes and compartments taken into consideration are defined in Figure 2-2. As the steady state fugacity model assumed, each section in this river basin is completely mixed. For each section, an equilibrium and steady state model treats four compartments: air, soil, water and sediment. Steady-state equation is written in fugacity

format to simplify them. Fugacity (f , Pa) is a criterion of equilibrium of chemical partition between phases, which is related to the concentration (C , mol/m³) by the expression

$$C_i = Z_i f \quad (1)$$

Where Z_i (mol/ m³.K) is the fugacity capacity of chemical in compartment. The function of Z value depends on the physical and chemical properties of the chemical and various characteristics of each compartment. Table 2-1 shows the fugacity capacity formulation of each compartment.

Whereas, f_i is fugacity (Pa) in compartment i . Fugacity (f , Pa) represents the tendency of the fluid to escape or expand isothermally and has units pressure (Mackay, 2001). For level II fugacity model, there is no active transport between environmental media. This assumption means that a single fugacity exists in the environment (Equation 2). Therefore, the fugacity of a chemical in a homogenous compartment as shown in Equation 2;

$$f_{AIR} = f_{SOIL} = f_{WATER} = f_{SEDIMENT} = f_i = f \quad (2)$$

Fugacity, f can be derived as;

$$f = \Sigma C_i V_i / \Sigma V_i Z_i \quad (3)$$

Thus, fugacity, f ;

$$f = M / \Sigma V Z \quad (4)$$

Where f is fugacity (Pa), M is total amount of chemical (mol) and V is volume (m³) of a compartment and Z is fugacity capacity (mol/m³.Pa) of chemical in a compartment.

Table 2-1 Fugacity capacity formulation

Compartment	Symbol	Fugacity capacity
Air	Z_{AIR}	$1/RT$
Soil	Z_{SOIL}	$\Phi \text{OC}_{\text{SOIL}} \times Z_{\text{SOIL}} \times \rho_{\text{SOIL}} \times \text{KOC}/1000$
Water	Z_{WATER}	$1/H$
Sediment	Z_{SEDIMENT}	$\Phi \text{OC}_{\text{SEDIMENT}} \times Z_{\text{SEDIMENT}} \times \rho_{\text{SEDIMENT}} \times \text{KOC}/1000$

Note:

R : Gas constant ($8.13 \text{ Pa}\cdot\text{m}^3\text{mol}^{-1}\text{K}^{-1}$)

T : Ambient temperature (K)

H : Henry's law constant ($\text{Pa}\cdot\text{m}^3/\text{mol}$)

$\Phi \text{OC}_{\text{SOIL}}$: Organic content in soil (-)

ρ_{SOIL} : Density of soil (kg m^{-3})

KOC : Octanol-water partition coefficient of chemical

Z_i : Fugacity capacity of compartment i ($\text{mol}/\text{m}^3\cdot\text{Pa}$)

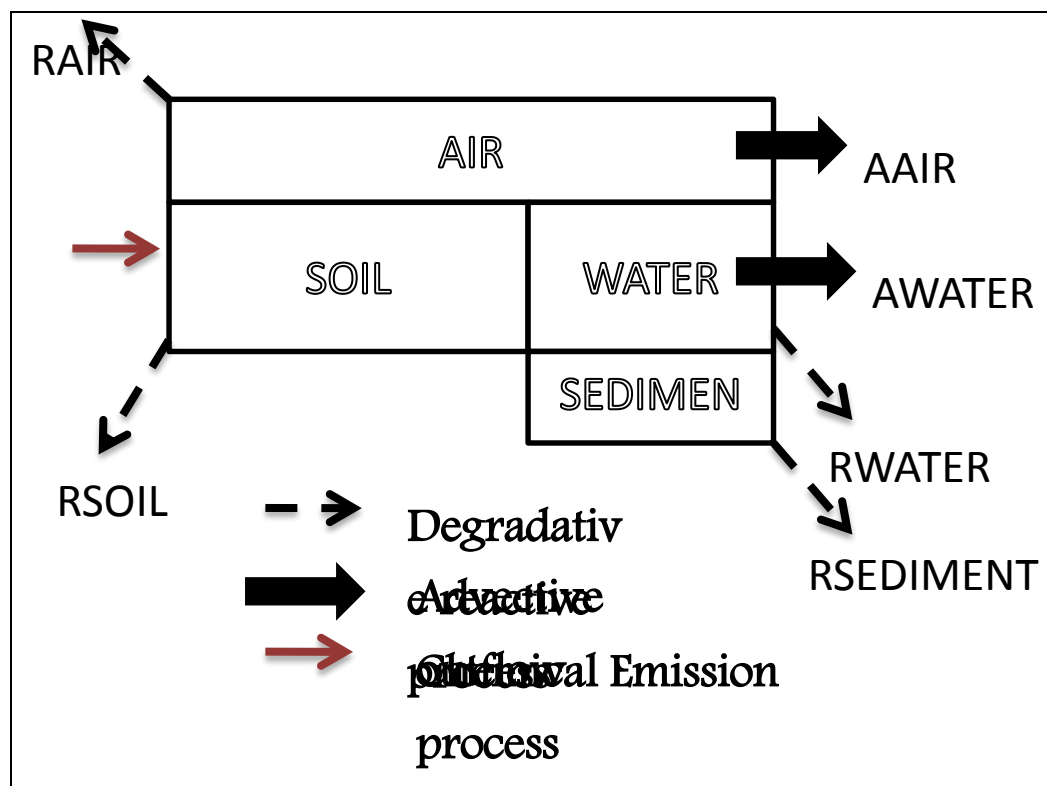


Figure 2-2 Compartments and processes

Besides, the transfer fluxes that describe the chemical movement in each compartment can be calculated based on transfer coefficient (D , mol/h.Pa) and fugacity based on Mackay, 2001. Under a steady state assumption, the process taken into consideration included advective air and water flow and degradation in air, soil, water and sediment. These transfer processes and corresponding D value are summarized in Table 1.

The general mass balance equation was established in term of transfer fluxes in each phase.

$$I = E + \sum D_{Ai} + \sum D_{Ri} \quad (5)$$

Where, I is total flux, E is total emission rate, D_{Ai} is advective fugacity rate constant in compartment I and D_{Ri} is degradation fugacity rate constant in compartment i .

Table 2-2 Transfer processes and D value

Symbol	Transfer process	Related D value
A_{AIR}	Advection in air	$(V/t)_{AIR} Z_{AIR}$
D_{AIR}	Degradation in air	$V_{AIR} Z_{AIR} k_{AIR}$
D_{SOIL}	Degradation in soil	$V_{SOIL} Z_{SOIL} k_{SOIL}$
A_{WATER}	Advection in water	$(V/t)_{WATER} Z_{WATER}$
D_{WATER}	Degradation in water	$V_{WATER} Z_{WATER} k_{WATER}$
$D_{SEDIMENT}$	Degradation in sediment	$V_{SEDIMENT} Z_{SEDIMENT} k_{SEDIMENT}$

Note:

V_i : Volume of compartment (m^3)

t_i : Residence time (h)

Z_i : Fugacity capacity of chemical in compartment ($mol/m^3.Pa$)

k_i : Reaction half- life in compartment(h)

k_i : Reaction half -life in compartment (h)

2.4 Model calculation

After the collection of related parameters of target chemicals and the river basin were achieved, the concentrations and transfer flux of DDT in air, soil, water and sediment at 3 different sections were calculated and coded using Visual Basic Application (VBA) from Microsoft Excel. Then, the calculated results were compared with measured results to validate the model.

2.5 Input Parameters

Parameters including physical–chemical properties of DDT, environmental parameters, and emission rate were collected for the model. The physical–chemical properties of DDT derived from Mackay *et al.* (2006) were used as input of this model as presented in Table 2-3. Environmental parameters used in this model such as compartment properties, organic carbon contents and others. Compartment properties such as compartment area and height obtained from landused map as spatial analysis using GIS software. Table 2-4 shows the area of landused type of each section. The area of compartment was estimated from latest land used map which provided from Johor Government Databases. The area of air was assumed as the total area of each section, whereas the area of sediment was followed exactly the area of surface water. While the height of each compartment based on several literature and study site value was listed in Table 2-5 . Other compartment properties listed in Table 2-6. DDT concentration was based on average concentration of DDT of each section from Department of Agriculture, Johor. Concentration was then conversed into total emission by multiplying with landused of palm oil and agriculture of each section and volume of spraying per square meter (0.001 L/m²). Once a week result was applied as spraying frequency. Table 2-7 shows estimated DDT emission rate were obtained from Department of Agricultural and site-survey.

Table 2-3 Physical-chemical properties of DDT

Properties	Symbol	Unit	Value
Molar mass	M	g/mol	354.49
Gas constant	R	Pam ³ /molK	8.314
Temperature	T	K	298
Vapour pressure	P _s	Pam ³ /molK	0.00002
Water solubility	C _s	mol/m ³	0.00012
Water solubility		g/m ³	0.0055
Henry Law Constant	H	Pam ³ /molK	0.166666667
Log octanol-water partition coefficient	LOGKOW	-	6.19
Octanol-water partition coefficient	KOW	-	1548816.619
	KOC	-	635014.8138
Reaction half-life in Air	t ^{1/2}	H	170
Reaction half-life Soil	t ^{1/2}	H	17000
Reaction half-life Water	t ^{1/2}	H	5500
Reaction half-life Sediment	t ^{1/2}	H	55000

- Physical-chemical properties obtained from Mackay,2006.

Table 2-4 Landused of each section (km²)

	Upstream	Midstream	Downstream
Total area	2060.14	17753.1	28202.68
Water	124.18	94.93	71.11
Transportation	205.64	76.77	144.73
Orchad	314.57	205.6	0
Palm oil	14535.79	15599.9	27565.29
Urban area	131.87	306.72	1003.25
Grooves	58.52	0	196.36
Mangroves	106.43	0.1	52.73
Vacant land	267.16	15.62	0
Rubber	47.68	0	284.59
Forest	3010.75	33.38	610.63
Agriculture	142.08	0	0

- Landused data provided by Johor Government Databases.

Table 2-5 Compartment height assumption (m)

Compartment	Upstream	Midstream	Downstream	Reference
Air	700	700	700	(Luo <i>et al.</i> , 2007)
Soil	0.2	0.2	0.2	(Luo <i>et al.</i> , 2007)

Water	4	4	3	Average assumption from Hydrology Department, UTM.
Sediment	0.05	0.05	0.05	(Luo <i>et al.</i> , 2007)

Table 2-6 Compartment properties

Properties	Symbol	Unit	Upstream	Midstream	Downstream	Reference
Organic carbon content in soil	$\Phi_{OC_{SOIL}}$	-	0.029	0.029	0.029	(Coulibaly, Labib, & Hazen, 2004)
Organic carbon content in sediment	$\Phi_{OC_{SEDIMENT}}$	-	0.025	0.025	0.025	(Coulibaly <i>et al.</i> , 2004)
Advective residence time for air	T_{AIR}	H	1	1	1	Mackay,2001
Advective residence time for water	T_{WATER}	H	240	240	240	Mackay,2001
Density of soil	ρ_{SOIL}	Kg/m ³	2400	2400	2400	(Luo <i>et al.</i> , 2007)
Density of sediment	ρ_{SOIL}	Kg/m ³	2000	2000	2000	(Luo <i>et al.</i> , 2007)

Table 2-7 Estimated DDT emission (mol/h)

Section	Emission rate
Upstream	0.0056
Midstream	0.1975
Downstream	0.4231

- Emission rate was obtained from the calculation of average concentration of DDT and landused at each section. The average concentration was obtained from Department of Agriculture, Johor.

2.6 Sensitivity Analysis.

A good overview of the most sensitive components of the model can be determined through sensitivity analysis. This analysis provides a measure of the sensitivity of parameters, forcing functions or sub-models to the state variables of greatest interest in the model. In practical modeling, the sensitivity analysis is carried out by changing the parameters, forcing functions and sub-models, and the corresponding response of the selected state variables is observed (Jorgensen, 1994). In this study, the sensitivity analysis was performed only for the parameters. A change for the parameter at $\pm 10\%$ was chosen, and the sensitivity coefficient (S) was calculated by the following formula (Jorgensen *et al.*, 1994):

$$S = (Y_{1.1} - Y_{0.9}) / (0.2 \times Y) \quad (6)$$

The terms, $Y_{1.1}$ and $Y_{0.9}$, represent the estimated concentrations when the tested parameter was changed at $+10\%$ and -10% , respectively. The greater the absolute value of sensitivity coefficient, the more sensitive the parameter.(cite)

3. RESULT AND DISCUSSION

3.1 Modeled Concentration of DDT

Modeled concentration of DDT in air, soil, water and sediment at upstream, midstream and downstream areas of Sayong River Watershed were obtained from modeled calculation carried out under the steady state, equilibrium and homogeneous assumptions. The calculation based on fugacity approach. The calculated DDT concentration in air, soil, water and sediment are presented in Figure 3-1.

Based on the modeled results, DDT mostly found in soil compartment followed by sediment. At upstream area, soil has the highest concentration of DDT with $1.506 \times 10^{-5} \text{ mol/m}^3$ followed by sediment with $1.082 \times 10^{-5} \text{ mol/m}^3$ and water $3.408 \times 10^{-10} \text{ mol/m}^3$ and least in air with $1.770 \times 10^{-13} \text{ mol/m}^3$ as can be seen in Figure 3-1 (a) . Next, Figure 3-1 (b) shows that in midstream, DDT mostly found in soil ($1.213 \times 10^{-4} \text{ mol/m}^3$) followed by sediment ($8.716 \times 10^{-5} \text{ mol/m}^3$), water ($2.745 \times 10^{-9} \text{ mol/m}^3$) and air ($1.428 \times 10^{-12} \text{ mol/m}^3$). Whereas for downstream area DDT mostly highest in soil ($1.744 \times 10^{-4} \text{ mol/m}^3$) followed by sediment ($1.253 \times 10^{-4} \text{ mol/m}^3$), water ($3.946 \times 10^{-9} \text{ mol/m}^3$) and air ($2.053 \times 10^{-12} \text{ mol/m}^3$) as shown in Figure 3-1 (c).

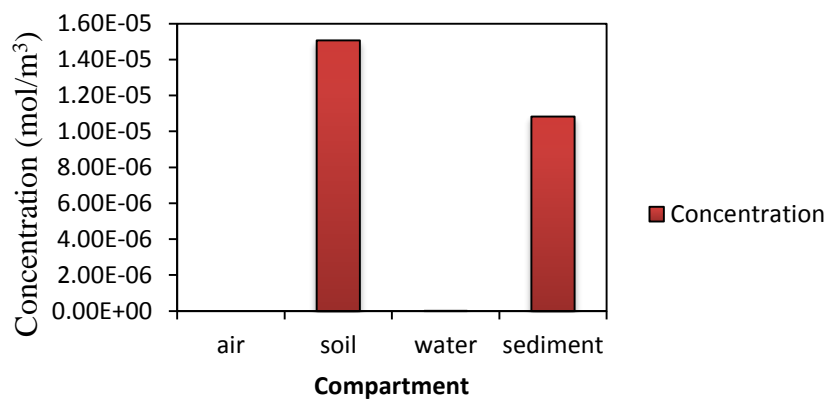
In this study, only soil compartment was assumed to receive direct emission source of DDT. Therefore, most DDT was distributed in soil. Sediment received no direct DDT or advective input. However, the accumulation of DDT might caused by soil erosion process.

Previously, the distribution of OCP which was hexachlorocyclohexana (HCH) in water, air and sediment at Lake Chaohu, China were simulated with a developed fugacity-based. Based on

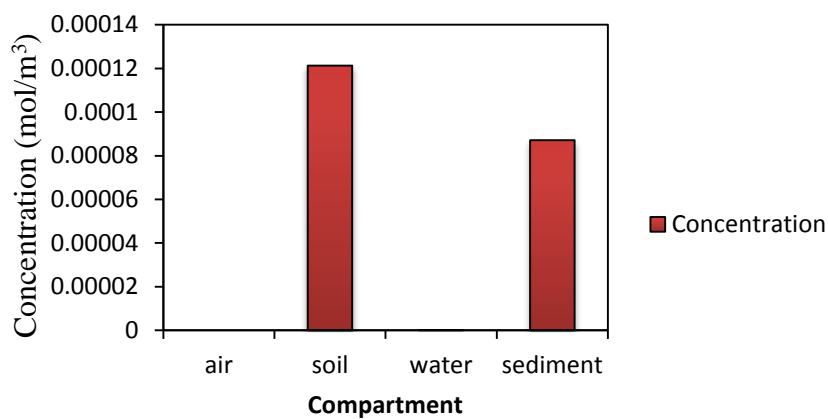
1 the study HCH was found highest in sediment phase ($1.00 \times 10^{-4} \text{ mol/m}^3$) followed by water (1.05
2 $\times 10^{-6} \text{ mol/m}^3$) and air ($1.07 \times 10^{-11} \text{ mol/m}^3$) (Kong *et al*, 2014). Besides, the simulation on
3 triclosan (TCS) and triclocarban (TCC) distribution in the Dongjiang River Basin, South China
4 was conducted using fugacity concept. According to the study, same to OCPs, compound of TCS
5 and TCC were highest in sediment $4.162 \times 10^{-5} \text{ mol/m}^3$ and $3.82 \times 10^{-4} \text{ mol/m}^3$, respectively,
6 followed by soil $5.31 \times 10^{-11} \text{ mol/m}^3$ and $1.29 \times 10^{-14} \text{ mol/m}^3$ for TCS and TCC, respectively.
7 (Zhang *et al*, 2013).

8
9 If compared for three sections (upstream, midstream and downstream), DDT in soil and
10 sediment at downstream area was highest followed by midstream and upstream. This happened
11 because the emission rate of DDT at downstream area was higher compared to midstream and
12 upstream.

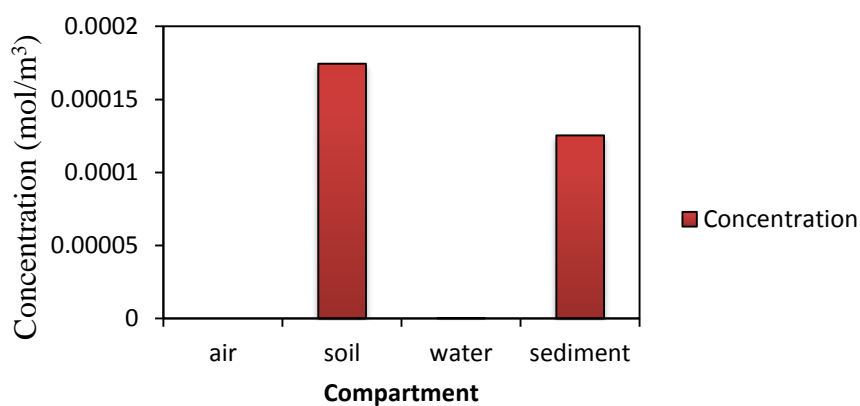
13
14 In short, it can be seen from the model, concentration of DDT in soils were highest for all
15 sections (Figure 3-2). This is because reaction half-life ($t^{1/2}$) of DDT in soil and sediment were
16 higher than air and water. Thus, DDT degraded slowly and attached longer at the soil and sediment
17 particles . The dissolved DDT concentration in the water was lower with the increase of their
18 molecular weight, probably due to the decreasing solubility.



(a)



(b)



(c)

Figure 3-1: Modeled concentration of DDT at Sayong River watershed for (a) upstream (b)midstream and (c)downstream

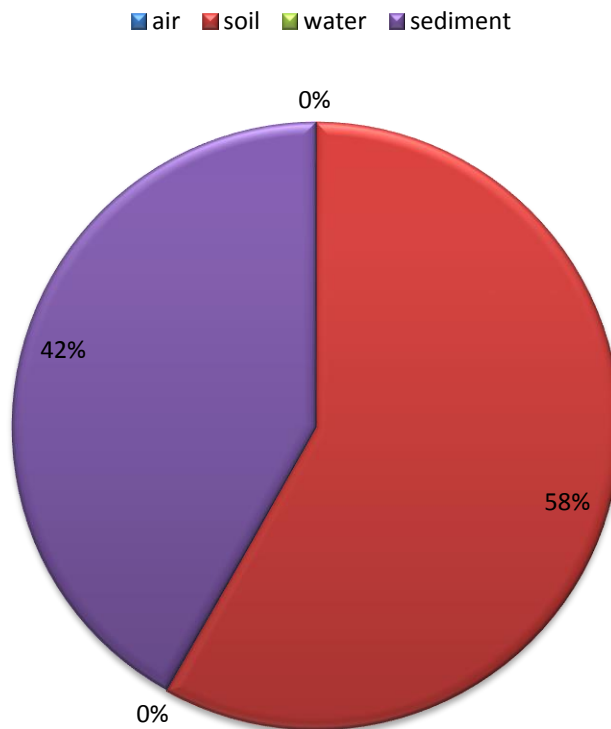


Figure 3-2: Overall percentages of DDT in air, soil, water and sediment for upstream, midstream and downstream of Sayong River watershed

3.2 Comparison of DDT concentration with field data

In this study, Gas Chromatography-mass spectrometry (GC-MS) analysis showed that there was no dichlorodiphenyltrichloroethane (DDT) detected in most air samples collected at Sayong River Watershed except in midstream area. In midstream area, the mean concentration of dichlorodiphenyltrichloroethane (DDT) in air was 0.37 µg/g (Figure 3-3) which was lower than soil and sediment. The presence of this compound in air was supported in previous study. For instance on 2011, a study of DDT was conducted to measure its distribution in atmosphere.

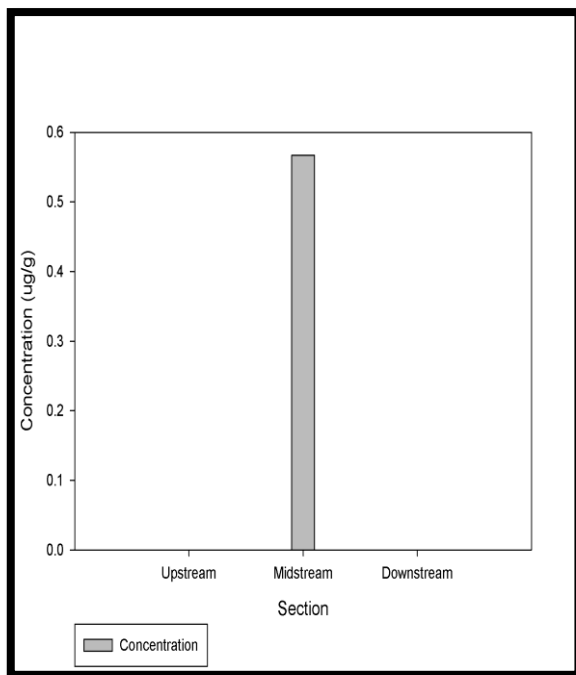
dichlorodiphenyltrichloroethane (DDT) was also detected in air samples. For examples, dichlorodiphenyltrichloroethane (DDT) was found at metropolitan City, Turkey such as Mudanya Coastal (22.5 pg/m³), Butal Traffic (41.9 pg/m³), UUC semi- rural (12.5 pg/m³) and Yavuz Selim Urban (49.5 pg/m³)(Cindoruk, 2011). The possible source of dichlorodiphenyltrichloroethane (DDT) in the air because of agricultural and vector elimination purposes. For instance, the usages lead to high amount of DDT detected in air at the tropical coastal atmosphere in India (Ali *et al.*, 2014).

Next, total dichlorodiphenyltrichloroethane (DDT) concentrations in soil of the Sayong River Watershed was detected in all soil samples. It can be seen from the Table 4.1 and Figure 4.1 (b), DDT in soil found highest in downstream area with the mean concentration 37.8763 ug/g followed by midstream (17.5020ug/g) and upstream with 0.2197 ug/g. This might happened because downstream area mostly covered with palm oil agriculture with the area of 27565.29 m² followed by midstream 15599.9 m² and downstream 14535.79 m². Thus, the use of DDT at downstream area might be high. Previously, the DDTs (the sum of P,P0-DDE, P,P0-DDT and O,P0-DDT + P,P0-DDD) concentration in soil samples were also detected along Chao River, China ranged from 0.1835 to 15.7150 ng g⁻¹(Yu *et al.*, 2014)

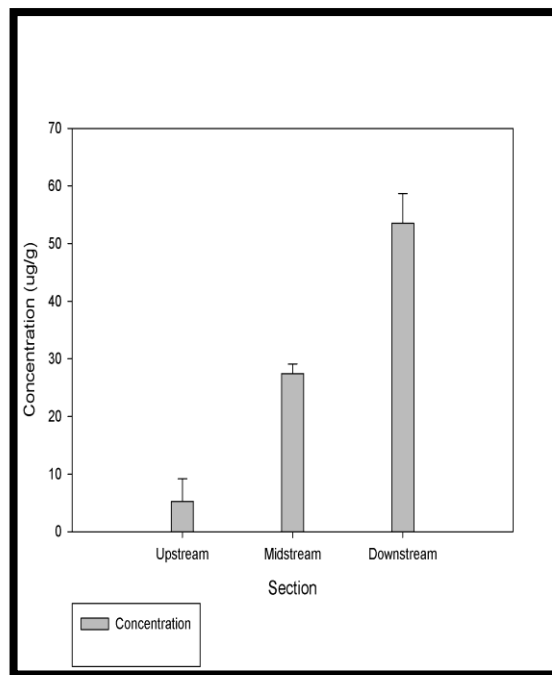
Besides, the total DDT concentrations in sediment of the Sayong River Watershed ranged from 0.3920 to 21.3 ug/g with a mean value of 10.0 ug/g. DDT was detected in most sediment samples. It can be seen from Figure 3-3 (c), DDT in sediment found highest in downstream area with the mean concentration 37.8763 ug/g followed by midstream (17.5020ug/g) and upstream

with 0.2197 ug/g. This might happened because downstream area mostly covered with palm oil agriculture with the area of 27565.29 m² followed by midstream 15599.9 m² and downstream 14535.79 m². Thus, the use of DDT at downstream area might be high.

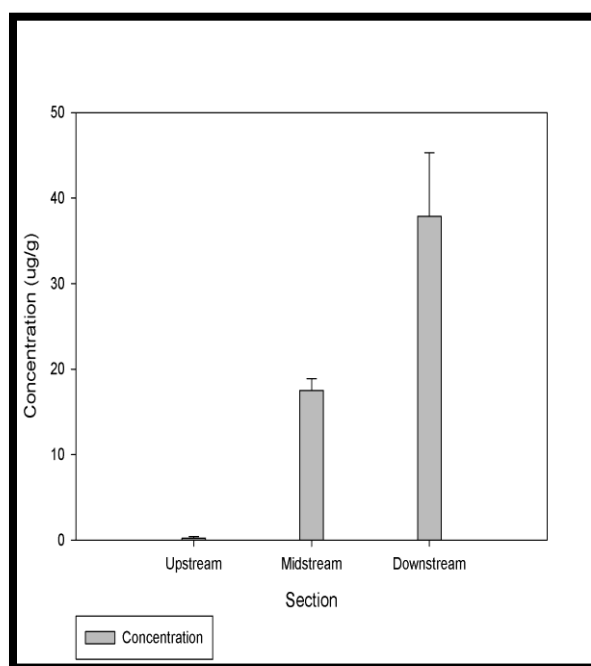
In this study there was no DDT detected in all collected water samples. If it was present, the concentration was very low. This might happened because most organochlorine pesticides (OCPs) have an affinity for particulate matter and one of their main sinks is marine sediments (Zhou, Zhu, Yang, & Chen, 2006).



(a)



(b)



(c)

Figure 3-3: Measured concentration of DDT in (a) air (b) soil and (c) sediment

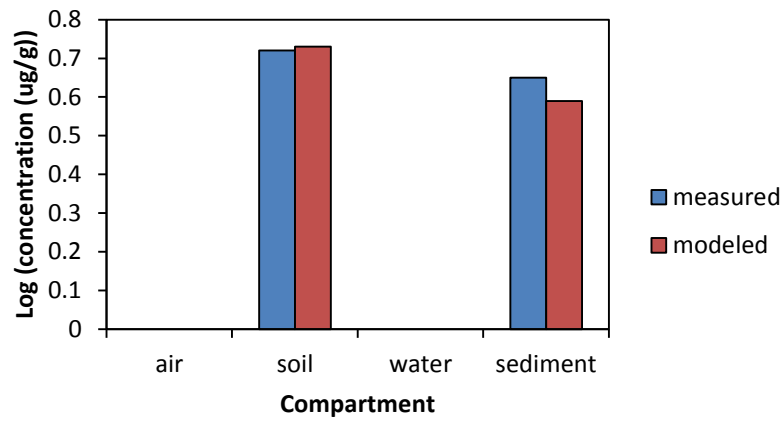
To test applicability of the developed multimedia model, modeling was performed for different sections of the Sayong River Watershed: upstream, midstream, downstream. The comparative results between the modeling data and monitoring data for the DDT compound are shown in Figure 3-4. The modeled concentration of DDT which was calculated in mol/m^3 was then converted into ng/g and ng/L . Then, both modeled and measured concentrations were converted into Log unit for comparison. As shown in the Figure 3-4, the modeled concentrations for DDT were comparable to and well matched with the monitoring data in all phases.

As shown in Figure 3-4 (a) the differences between the modeled and measured means in air, soil, water and sediment phases of the upstream area were 0.00, 0.01, 0.00 and -0.06 log units,

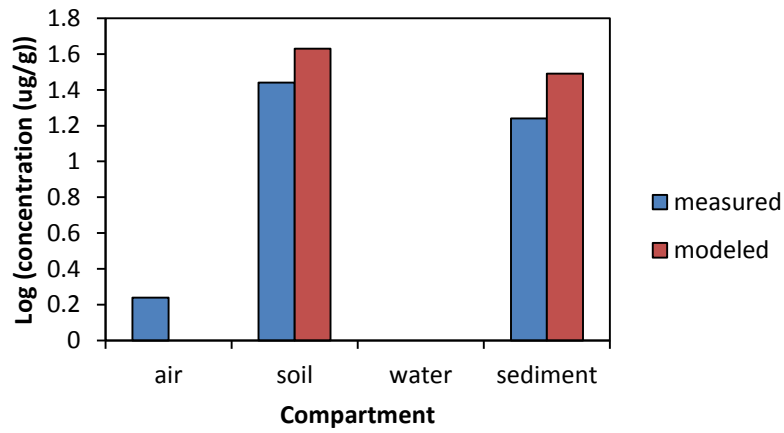
1 respectively. For midstream area, the differences between the modeled and measured mean were
2 -0.24, 0.19, 0.00 and 0.25 log units for air, soil, water and sediment, respectively (Figure 3-4(b)).
3 Whereas, the differences between the modeled and measured means in air, soil, water and sediment
4 phases of the downstream area were 0.00, 0.05, 0.00 and 0.02 log units, respectively as can be seen
5 in Figure 3-4 (c) (an acceptable range: 0.5–0.7 in log unit range).

6
7 For air, there were no differences between modeled and monitoring data for upstream and
8 downstream watershed. However, large residual DDT was found in midstream area (log different:
9 -0.24). In this case the modeled result was underestimated. Since OCPs do not have point sources,
10 they might enter the atmosphere by evaporation from previously contaminated soils, water bodies
11 and vegetation (Cindoruk, 2011). The difference occurred because in this study the model not
12 treated those processes. For water, there were no differences between modeled and monitoring
13 data for whole river watershed. This is the perfect matched of the modeled with 0.00 log different.
14 Besides, for soil the difference between modeled and monitoring data for whole watershed was
15 0.01-0.05 log units (in acceptable range). Measured concentrations were lower especially in
16 downstream area.

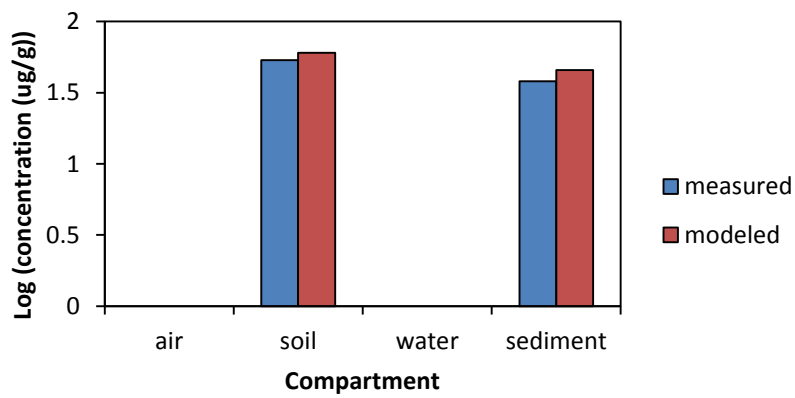
17
18 The differences between the modeled and measured DDT concentrations for air, soil and
19 sediment phases are attributable to the complexity of DDT source and the degree of influence by
20 environmental changes(Xu *et al.*, 2013). For all sections of the watershed, the obtained modeling
21 values agreed well with the monitoring observations with the residual within an order of
22 magnitude.



(a)



(b)



(c)

Figure 3-4: Comparison between measured and modeled concentration of DDT in air, soil, water and sediment in (a) upstream (b) midstream and (c) downstream of Sayong River watershed

3.2 Transfer Flux of DDT

To exhibit the spatial impact on the transfer processes, transfer fluxes for DDT in different sections (upstream, midstream and downstream) of the Sayong River watershed are presented in this study. This model addresses the effect of advection outflow and degradation reaction on DDT fate. The fluxes into and out of the watershed area were well-balanced. Based on Table 3-1, the relative error the total flux into and out of the upstream area was 0.0001 (mol/h), for midstream 0.0019 (mol/h) and downstream 0.0008 (mol/h).

1 **Table 3-1: The contribution of each transfer flux in and out of the four compartments**
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Section	Compartment	Process In*	Coefficient rate (mol/h)	Process Out*	Coefficient rate (mol/h)
Upstream	Air		-	A _{AIR}	0.25670
		Emission		R _{AIR}	0.00100
	Soil	(E)	0.5642	R _{SOIL}	0.30502
	Water		-	A _{WATER}	0.00055
				R _{WATER}	0.00002
	Sediment		-	R _{SEDIMENT}	0.00085
	Total (mol/h)		0.56242		0.5641
	Relative error			0.0001	
Midstream	Air		-	A _{AIR}	0.20285
		Emission		R _{AIR}	0.00083
	Soil	(E)	0.2257	R _{SOIL}	0.02146
	Water		-	A _{WATER}	0.00046
				R _{WATER}	0.00001
	Sediment		-	R _{SEDIMENT}	0.00006
	Total		0.2257		0.22385
	Relative error			0.0019	
Downstream	Air		-	A _{AIR}	0.12159

			R_{AIR}	0.00050
Soil	Emission	0.1269	R_{SOIL}	0.00455
Water	(E)	-	A_{WATER}	0.00028
			R_{WATER}	0.00000
Sediment		-	$R_{SEDIMENT}$	0.00002
Total (mol/h)		0.1269		0.12614
Relative error			0.0008	

*The in refer to input flux and the out refer to output fluxes. All the abbreviation can be referred in list of symbol/abbreviation

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Transfer fluxes of DDT at Sayong River watershed are presented in Figure 3-5 Based on the calculated result, the loss rate process of DDT through advective air flow out of the basin area (A_{AIR}) and degradative reactive in soil (R_{SOIL}) were predominant in the output of DDT. Based on figure 4-5, A_{AIR} and R_{SOIL} mostly occurred in upstream area (A_{AIR} (0.2567 mol/h) and R_{SOIL} (0.3050 mol/h)) followed by midstream (A_{AIR} (0.2029 mol/h) and D_{SOIL} (0.0215 mol/h)) and downstream (A_{AIR} (0.1216 mol/h) and D_{SOIL} (0.0045 mol/h). Among all the loss processes, degradative reactive in water (R_{WATER}) was least occurred for all sections with the rate 0.00002, 0.00001 and 0.00000 for upstream, midstream and downstream, respectively.

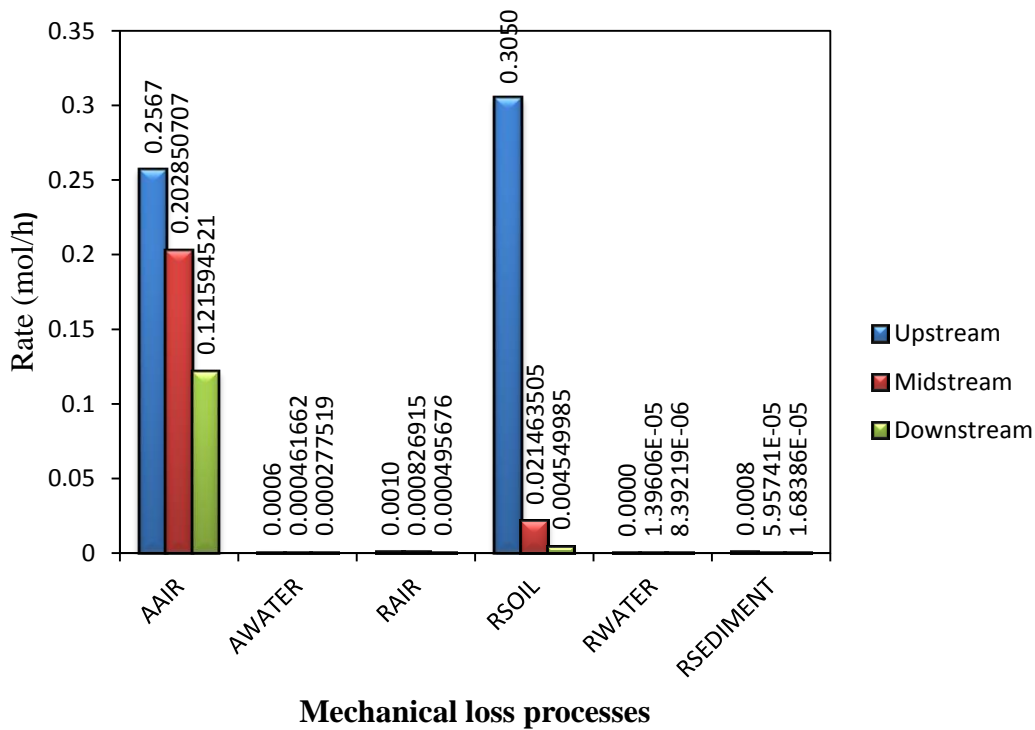


Figure 3-1 Modeled transfer fluxes of DDT at Sayong River watershed

3.3 Sensitivities of Modeled Concentration to Input Parameters

To compare the influence among parameters, a sensitivity coefficient (C_s) was adopted. The sensitivity coefficient higher than 0.5 (>0.5) are considered more influential parameter in the model. The model output included the concentration of DDT in four phases including air, soil, water and sediment in three sections; upstream, midstream and downstream. The results were summarized in Table 3-2. Among 21 input parameters, there were only 10 parameters with sensitivity coefficients higher than 0.5. Based on Table 5, Vapour pressure (P_s) and octanol-water partition coefficient (KOC) were the most influential parameters for upstream, midstream and downstream area in the model for DDT compound.

Table 3-2: Sensitivity coefficients of the more sensitive parameters for the model ($SC_i > 0.5$).

		T	Ps	Cs	KOC	K	ρ_d	ρ_p	Ocd	OCp	I
Upstream	Air	-	-3.1	-0.79	-6.7	6.6	-	-	-	-	1
	Soil	-	2.6	-	5.6	6.6	-	-	-	-	1
	Water	-	2.6	-	-6.7	6.6	-	-	-	-	1
	Sediment	-	2.6	-	5.6	6.6	-	0.9	-	0.9	1
Midstream	Air	-	-1.3	-	-1.5	1.18	-	-	-	-	1
	Soil	0.9	12.5	0.75	14.7	1.18	0.9	-	0.9	-	1
	Water	0.9	12.5	0.75	-1.5	1.18	-	-	-	-	1
	Sediment	0.9	12.5	0.75	14.7	1.18	-	0.9	-	0.9	1
Downstream	Air	-	-0.6	-	-0.7	-	-	-	-	-	1
	Soil	0.9	16.2	0.89	17.4	-	0.9	-	0.9	-	1
	Water	0.9	16.2	0.89	-0.7	-	-	-	-	-	1
	Sediment	0.9	16.2	0.89	17.4	-	-	0.9	-	0.9	1

4. CONCLUSION

A fugacity concept was applied to develop a multimedia environmental fate and transport model of DDT at the Sayong River watershed. By using the developed model, the calculated concentrations of four DDT in various environmental media (air, soil, water and sediment) were obtained. The DDT was found to have a tendency to be distributed in soil and sediment for all sections (upstream, midstream and downstream). Transfer flux analysis indicates that advective air flow was the primary process for the disappearance of the chemicals in the area, followed by degradative reactive process. Besides, vapour pressure (P_s) of the chemicals, octanol-water partition coefficient (KOC), water solubility (C_s) and absolute temperature (T) were some key influencing parameters for chemical distribution in the environment. When compared with field data, the modeled concentration of DDT was in good agreement with measured concentration. As suggestion this steady state fugacity model can be used to evaluate contamination of long range transport chemicals at the basin scale based on their usage. This tool could be adopted by various decision-makers in the management of chemicals at Sayong River basin.

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