

Risk Assessment on Landscape Development using Remote Sensing and Geographic Information System (GIS): Assessment of the Impact of Agricultural Development on Soil Loss and Degradation of Ecological Service Values and Goods

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INTRODUCTION

Agriculture plays an important part in the overall economic development of most developing countries like Malaysia through its contribution to the Gross Domestic Product (GDP), employment and foreign exchange earnings. In 1996, the share of agriculture has declined to 13.6 percent. Total agricultural production for the period from 1994 to early 2000 showed an increasing trend. In the developing world, the need to provide adequate food for a rapidly expanding population has put enormous pressure on existing productive land and brought marginal land under cultivation. This has result in deforestation of the tropical forests and drainage of wetland areas. In semi arid and arid regions, lands have been brought into production with use of irrigation. In many cases, this has led to an accumulation of salt in the soil (salinisation). The subsequent loss of vegetation cover leaves the land open to soil erosion and eventual desertification. Soil erosion represents one of the most serious depletions of natural resources in the world today. Rapid forest conversion and land development activities have resulted in accelerated soil erosion in many parts of the word. Global erosion rate are estimated at 75×10^9 ton/year (Tan & Shibasaki, 2003). Over the past several thousand years, deforestation, crop agriculture and grazing have promoted the greatest soil erosion and these activities continue to hold that dubious distinction on a worldwide basis today.

Apart from the above erosion effect, the accumulated effect to agricultural development is degradation of ecology service values and goods (ESVG) not only in the area in which the development taken place but also effecting the linking ecosystem services. Ecosystem services have been estimated (Constanza *et al.*, 1997) for realism of guideline for evaluating ESGV prior and after any land related development were taken place. Using Pasoh Forest Reserve (PFR) as study area, this study is a part of joint research between Japan and Malaysia scientists in understanding various issue related to tropical forest ecosystem (NIES, 2003). We collect, analyze, and deliver vital ecological information to decision makers pertaining to forest-related industries and management. Most importantly these information are conveyed in simple form, easy to understand together with the economic values for any decision on developing the land or changing it use. To meet management needs, the landscape models are required in assess the effects of different management scenarios. These scenarios often require decision-making horizons spanning broad temporal and spatial scales. Often, simulation models are the only way to assess alternatives that could be tested under such real-world conditions (Mladenoff & Baker, 1999). From this synthesis, the value for ecosystem service per unit area for each landscape feature can, therefore be estimated. Although there are problems inherent in producing such an estimate (Blamford *et al.*, 2002), this mechanism is essential in order to make the range of potential values of the services of ecosystems more apparent.

The minimal impact risk attributed due to landscape development to its ecological regime has been the main concern in supporting sustainable development. However, in the tropical regions, risk assessment on landscape development still perceived to have minimum impact, judging from the rate of conversion of forested areas to other uses (Pulzl & Rametsteiner 2002, Adger *et al.* 1995, Bishop 1999, Peters *et al.* 1989). Conversion of logged-over forests to agricultural land, namely large scale plantation for industrial crop such as such oil palm is

perceive a good option in Malaysia economically other than creating better employment opportunities compared to the productive forest. This is economically true, evident by the amount of increasing standard of living to rural settlers of large scale plantation schemes in Malaysia since early 70's. Since then, conversion of forested land to agricultural activities particularly large-scale oil palm plantation area has profoundly increased from 320 thousands to 2.1 millions hectares in 2000; and it is projected to occupies 5.10 millions hectares by 2020 (Jalani *et al.*, 2002). In 2000, about half of rubber plantations opened since 1970 was converted to oil palm due more profitable returns of export commodity prices of palm oil products.

In economical term, the market value of a land is determined by its size, location and the accessible infrastructural facilities. The likelihood of land use change in the vicinity also influences the current land value, i.e. maintaining a forest reserve within designated residential area is favorable to the current and future value and vice-versa. Apart from economic value, the ecological service values to any piece of land area that are not at all accounted into the present market values. Worst still these ecological values are evaluated based on scientific method to benefits human (Constanza *et al.*, 1997; de Groot, 1992; de Groot, 1987) not widely known to political masters, decision makers and even majority of the stakeholders in tropical regions, where majority are found in the developing countries. In these regions, sustainable economy and its growth seems to be the paramount agenda compared to sustainable development (ADB, 2004), although such a growth only achievable through sustainable land-related development. Subsequently, the mapping of ecology service values and goods (ESVG) of PFR for 1995 and 2003 is undertaken. Trends of changes in land use pattern within the temporal data sets and ESGV were then analyzed and the impacts were then determined.

MATERIALS

Study area

The present study was conducted in an old-growth, lowland dipterocarp forest within the Pasoh Forest Reserve (lat 2°59'N, long 102°18'E), which is located in the state of Negeri Sembilan, about 70 km southeast of Kuala Lumpur, Malaysia (see Figure 1). The mean annual rainfall from 1974 to 1992 at Pasoh-Dua (lat 2°56'N, long 102°18'E), 6 km south of the reserve, was 1842 mm, with distinct rainfall peaks in April-May and November-December (based on data provided by the Malaysian Meteorological Service). The soil is the Bungor-Malacca Association Type (based on data provided by the Malaysian Soil Science Division), which develops mainly from shale, granite, and fluvial granite alluvium parent materials (Allbrook, 1973). The topography consists mainly of flat alluvial areas, with smaller expanses of swales, riverine areas, and gently rolling hills.

The overall vegetation type in the reserve is lowland dipterocarp forest, which is characterized by a high proportion of species in the Dipterocarpaceae (Okuda, 2003). On the basis of floristic evidence, the primary forest in the study area was generally homogeneous, with no evidence of major disturbance, and appeared to be representative of the lowland forest of the south-central Malay Peninsula. Lowland dipterocarp forest is one of the most species-rich communities in the world, with more than 200 tree species per hectare. In contrast, the southern and eastern edges of the reserve had been selectively logged from the mid-1950s until the early 1970s, and at the time of the study represented regenerating forest.

Ecological service values and goods

The ecological service values and goods are defined by Constanza *et al.* (1997) as the benefits human populations derive directly or indirectly from the ecosystem functions. The ecosystem functions, on the other hand refer to various habitat, biological or system properties or processes of ecosystems. Both the ecological service and goods often combined for simplicity and termed as ecological services. Large numbers of ecosystem services have been identified and reported in many previous studies such as de Groot (1987), Turner (1988) and

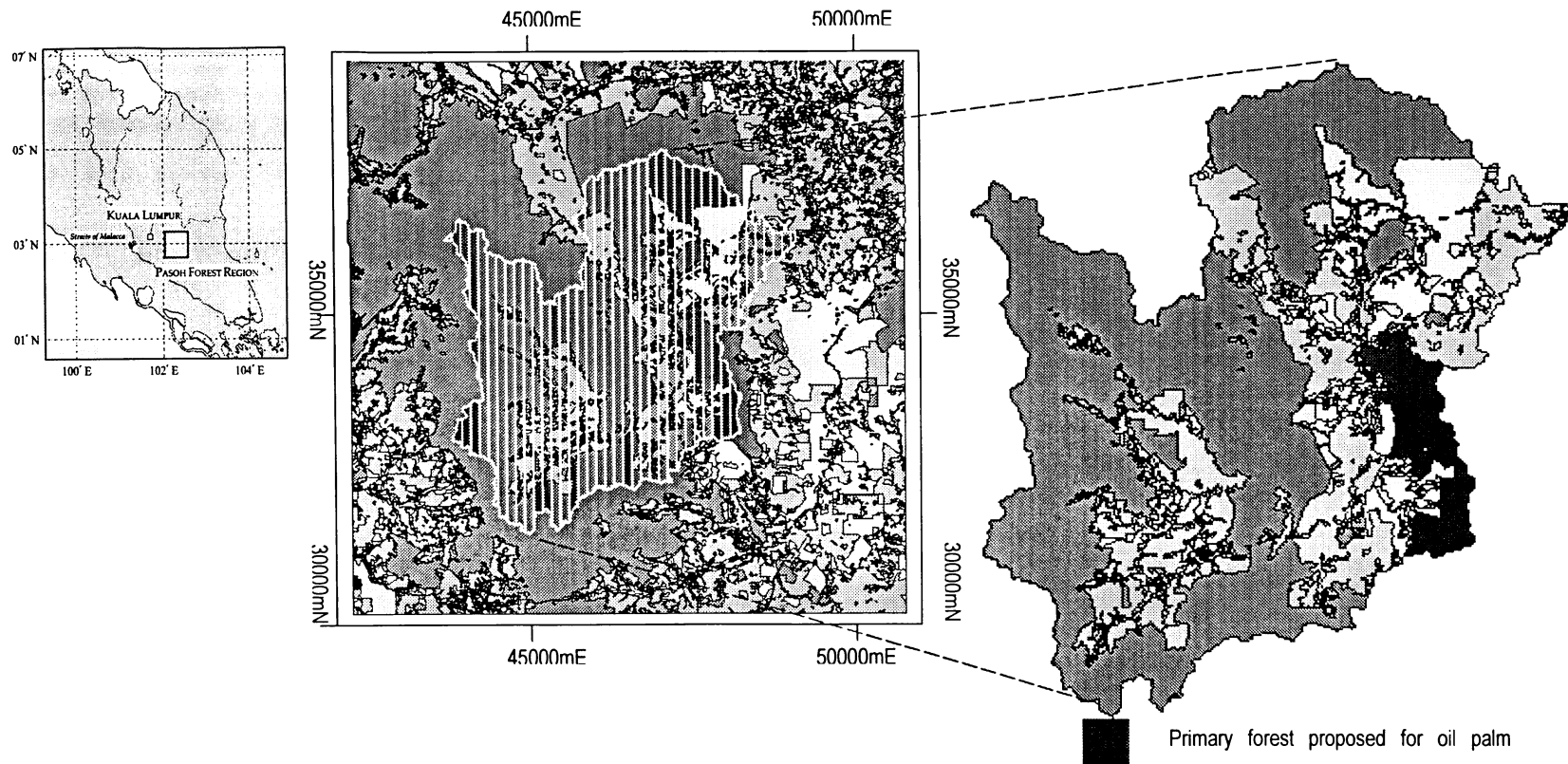


Fig. 1. Study area; Pasoh Forest Reserve and vicinity (PFR) with area of 0.8 millions hectares. The Sungai Teriang catchment used in simulation of the cost of annual ecological services values and goods against the asset generated by the proposed by converting existing about 339,630ha of primary forest (demarcated in solid red colour) into oil palm plantation

de Groot (1992). A total of 17 attributes of ESGV (Constanza *et al.*, 1999) are considered here within the context of tropical ecosystems. These categories only included renewable ecosystem services adopted from average global value of ESGV. The ecosystems services and functions do not necessarily show a one-to-one correspondence; hence, some of these ecosystem service values are still contentious, particularly in developing countries like Malaysia where the pressure of land related developments is so great for obtaining better standards of living. In this study, two sets of ESGV were computed for PFR. First, ESGV based on full 17 attributes of ESGV, and secondly selected 11 attributes only, customized to sounds logic to have direct relationship with the land use. There is large variation in the value ESGV over the years since it was incepted (early 90's), because such values are very much subjective to large variety of countries location and priorities in their environmental up keeping status. ESGV for examples are well established and highly priced in developed nations but in the rest of world, even the ESGV contention seems unnecessary when there are more prioritized critical agendas. By using these values, it formed as attributes to the corresponding spatial data, readily created in spatial database, thus mapping of ESGV is possible.

METHODS

The entire method employed in this study is divided into three: (i) analysis of impact of agricultural development on soil loss from 1995 to 2003, (ii) mapping and analysis of changes in the ecology services values and goods in 1995 and 2003, and (iii) simulation of impact on ecology services values and goods for period of five years for converting secondary forest into palm oil plantation within a catchment in the study area. The followings describe each of the method employed in the analysis.

Analysis of impact of agricultural development on soil loss from 1995 to 2003

The total soil loss of PFR was determined using universal soil loss equation (USLE) (Morgan, 1974) for 1995 and 2003 was used to analyze the impact of agricultural development on soil loss. The soil loss in USLE is calculated, such that:

$$A = R.K.L.S.CP \quad (1)$$

where A is soil loss (ton/ha/yr), R is rainfall erosivity factor, K is soil erodibility factor, L is slope length factor, S is slope-steepness, C is crop management factor and P is erosion control practice factor.

The land use pattern for both 1995 and 2003 were determined using multi-temporal Landsat-5 Thematic Mapper data set, were used as C -factor in USLE. The CP factor can be replaced by vegetation management factor, and for Peninsular Malaysia the values of CP classes found in PFR is tabulated in Table 1. The erosivity factor K used is also tabulated in Table 1. Within an area, the total soil loss calculated per set given land use pattern in 1995 and 2003, hence, can be therefore use as basis for landscape risk assessment of agricultural development on soil loss.

Mapping and analysis of changes in the ecology services values and goods

Using GIS the land use maps (1995 and 2003) produced by classifying multi-temporal Landsat TM data sets and other ancillary data were compiled in spatial database of PFR, from where the ecology service values were derived. These values referred to certain ecological system values and goods per given area. The absolute ecological service values and ecology goods (ESVG) are very critical in making the results realistic as in real world. As guidelines to such approach, adaptation to special paper released by FAO (Bann, 1998) is used in this study. The ecological services focused for PRF are those related only tropical forest landscape using partial (11

Table 1 The C and P factors found in PFR for simulating the risk of landscape management options (a), and the K factor used (b)

(a)

Land use options	C	P	C*P
1. Agriculture			
• Oil Palm plantation	0.5	0.25	0.125
• Rubber plantation	0.2	0.25	0.100
• Orchard (inc sundry tree cultivation)	0.3	0.5	0.015
• Pasture	0.02	1	0.020
2. Forest			
• Primary Forest	-	-	0.003
• Commercial forest (secondary forest)	-	-	0.015
• Shrubs (<i>belucar</i>)	0.02	1	0.020
• Clearings due to logging	-	-	0.015
3. Built-up areas			
• Urban	0.005	1	0.005
• Housing	0.003	1	0.003
• Cleared/barren land	1	1	1

(source: Hashim, 2004)

(b)

Soil Series	Symbol	% clay	% silt	% fine sand	% organic matter	K
Reverine Alluvium Telemong	RVA-TMG	62	21	10	6.25	0.071
Rengam- Tampin	RGM-TPN	22.5	6.5	26.5	2.5	0.190
Bungor- Durian	BGR-DRN	26.5	21	46.5	1.6	0.180
Batu Anam Durian	BTM-DRN	37	24.5	34	1.515	0.290
Tampin- Rengam	TPN-RGM	22.5	6.5	26.5	2.495	0.190
Durian-Malacca	DRN-MCA	42.5	20	35	1	0.260
Bungor- Malacca	BGR-MCA	42	15	35.5	1.6	0.230
Durian-Tavy	DRN-TVY	34.5	30.5	31.5	0.5	0.350
Inland-Swamp-Local Alluvium	ISA-LAA					0.280

(source: Hashim, 2004)

attributes) and full 17 attributes for ecological services and goods (Constanza *et al.*, 1997). The modified proposed 11 ecological values are only have directly linked to the landscape features that can be tangibly accepted by stakeholders, such that comprised of 7 ecological service values and 4 ecological goods. For the entire PFR, 10 landscape features are used to derive their respective ESGV.

Analysis on impact on ecology services values and goods for period of five years for converting secondary forest into palm oil plantation within a watershed area

Changing the land use option would also change in the C (crop management factor) and P values (erosion control factor) from the previous land use. Resultant of change in CP factor would effect differently in the erosion risk effect, such that the amount of soil loss in primary forest is within very low risk (mean of 10 ton/year) could change to variable risk range moderately high to moderately very high risk (50-100 ton/year). To change the land use option, say for example from its present land use, i.e. from primary forest to oil palm plantation, the value of resultant C*P would shift to 0.125 from 0.003 (please refer Table 1) which means that tendency of erosion is about 42 times from its present land use.

In simulating the landscape scenario changes, the ESGV on the existing land use and after the conversion is determined in an annual basis for period of ten years. In this study, an emulation of converting primary forest for large oil palm plantation. Figure 1, show the proposed primary forest (with an area of 339,630ha) for this project is fragmented forest, surrounded by large rubber and oil palm plantations apart from mixed sundry cultivations.

The critical issues in the creating the simulation are the costs for each ESGV for specific landscape features. Adaptation to ESGV proposed in annual world ecosystems (Constanza *et al.*, 1997) have used in calculating both these values in this study. FAO (2003) guidelines for estimates of economic values of forest benefits have also been adapted for complimenting the incomplete ecology goods (EG) in Contanza *et al.* (1997). The EG for oil palm plantation has been based on the average previous years of prices of oil palm products for 1995-2003 (MPOB, 2005) and special banker's report (Bank Bumiputra-Commerce, 2004). The average prices of oil products used in the EG are for prices of crude palm oil, palm kernel, crude palm kernel oil and palm kernel cake. Each plantation of 1ha produces an average of 18 ton of fresh fruit bunches a year. With average oil extraction rate is 20%, this yields an average of 3.5ton/ha/yr of crude palm oil. Analyzing the trend of production of oil palm products from 1995 to 2003, only 42.8% of the crude palm oil exported and the remaining are exported as oil palm products after further processed by downstream industries producing palm kernel (28.8%), crude palm kernel oil (12.9%) and palm kernel cake (15.5%). With this production ratio and the average yield generated from oil palm plantation is US\$2792/ha/yr and average cost incurred in the production is US\$1048/ha/yr. In addition, initial costs are; (i) acquiring and developing the land - US\$790/ha, (ii) waiting to crops maturity of first 3 years about US\$1711/ha/yr, (iii) upkeeping inclusive of fertilizer US\$252/ha/yr, and (iv) collection, general charges, manufacturing and dispatch of oil palm products at US\$353/ha/yr. Each of the breakdowns for the production costs have been based on Pow (2003). The detailed lists for production cost is similar to Moll (1987), however, the values are out of date.

RESULTS AND DISCUSSIONS

Agricultural crops formed 51.8 and 51.6 percents in 1995 and 2003 land use pattern, respectively. In PFR, the primary and secondary forests still occupy more than 44 percent of the land cover. Within 1995 to 2003, the changes in this landscape features is very minimal. The net change in the landscapes over 8-year period is the reduction of 1464ha (0.2 percent PFR area) of primary forest. Reciprocating deforestation, 12 other land use classes have increased in area - rubber, oil palm and mixed cultivation are major changes that have taken place. In total, the agricultural land development is taking place at annual rate of about 160ha per year. Table 2 tabulates

the land use pattern and its trend of changes within the 8-year period. Table 3 tabulates the soil erosion rate, impacts of this agricultural development. Within the vegetated land use classes, primary forest was found as expected remains the least affected by soil loss with mean soil loss about $10 \text{ ton ha}^{-1} \text{ yr}^{-1}$. Contrastingly, the mean soil loss in secondary forest is about $60 \text{ ton ha}^{-1} \text{ yr}^{-1}$, similar trend to that of rubber and oil palm plantations. The highest erosion rate is within mixed cultivation classes (trees and non-trees) with mean of 970 and 1071 ton/year in 1995 and 2003, respectively. Agricultural lands in PFR are contributing as main source of soil loss $53 \times 10^6 \text{ ton yr}^{-1}$ in 1995 and $63 \times 10^6 \text{ ton yr}^{-1}$ in 2003, i.e. 90 percent of total soil loss due to erosion in the PFR. Non-vegetated classes; the urban area, quarry and mines area contributed merely an average of less than 1 ton yr^{-1} . The mean soil loss in primary and secondary forests is comparable to studies in the Sg Gombak catchment (Douglas, 1968) in 40% forested area. The mean sediment yield of secondary forest in Berembun catchment (Baharuddin, 1988) show similar trend in the erosion rate of PFR, about 6 times of that primary forest. The erosion rate modeled by empirical analysis such using USLE as in this study is higher when compared to actual sediment yields record from in-situ measurement, however, both these values are relatively correlated, hence, widely used in assessing empirical erosion risk assessment. Please refer to Zulkifli (2005) for elaboration on this issue. The soil loss for each of the land use class in 1995 and 2003 were grouped into 4 erosion risk categories; namely: (i) very low risk ($0-1 \text{ ton ha}^{-1} \text{ yr}^{-1}$), (ii) low risk ($1-10 \text{ ton ha}^{-1} \text{ yr}^{-1}$), (iii) moderate risk ($10-50 \text{ ton ha}^{-1} \text{ yr}^{-1}$), and (iv) moderate high risk ($50-100 \text{ ton ha}^{-1} \text{ yr}^{-1}$). Table 3(b) tabulates the area extent for each class under each of these erosion risk categories. The net effect of erosion risk in PFR is low; only 0.1% of the area is under moderate to moderately high risk. The oil palm and rubber plantations posed the most erosion risks at various levels of growth. This is evident by the existing of erosion risk of very low to moderately high risk for these two crops, with majority of the soil loss within these two classes in $1-10 \text{ ton ha}^{-1} \text{ yr}^{-1}$.

Changes in land use patterns also contributes to the decreased of ecological services and goods for PFR, i.e. cost $\text{US\$}179 \times 10^6$ and $\text{US\$}114 \times 10^6$ in 1995 and 2003, respectively. Both these values accounted for the corresponding soil loss due to erosion at $\text{US\$}6/\text{ton}$. Table 4 and 5 tabulate the ecological service values and goods for PFR in 1995 and 2003. The decrease of $\text{US\$}65$ millions between the 8 year period is attributed to the conversion of 1389 ha (about 0.2%) of forested areas into agricultural development activities, mainly oil palm and rubber plantations apart from increasing importance of mixed sundry cultivation for cash crops. Suppose the conversion trend continues annually, the ESGV will deficit at about $\text{US\$}174 \text{ ha yr}^{-1}$. The question remains unanswered to most stakeholders and the government, where and how are the physical earnings from ecological services can generate income to sustain livelihood of population in the area? Until this issue of compensation for preserving ecological service values is ‘settled’, the $\text{US\$}174 \text{ ha yr}^{-1}$ conversion rate of PFR forested lands to other economical oriented classes is seem unstoppable to various pressures of sustaining better standard of living. Studies by Arif and Tengku Mohd (2001) indicated that poverty in oil palm smallholders reduce from 30.3 to less than 8% percents in 1970 and 1990 compared to other rural economic activities viz. rubber smallholders, coconut smallholders, paddy farmers, fishermen, estate workers and other agriculture. Mass plantation of commodity crops - rubber and oil palm have been initiated as early 1957 in Malaysia to resettle rural population to large plantation schemes with an objective to uplift their economics. Aggressive efforts of such schemes have met objectives and by 2000 with total cumulative area planted with oil palm amounted to 685,520ha. Most all agricultural land development is economically driven while ESGV is science-driven, too conceptual to sustaining environment in most developing countries that need the resources for development, particularly on virtual cost of ESGV.

Conversion of 339,630ha of primary forest to oil palm plantation would give significant impact to ESGV. Figure 2 illustrates the net value generated from keeping primary forest for 20 years against the asset derived from oil palm plantation. Considering 11 ESGV attributes, the entire annual ESGV values generated from oil palm plantation (maturity period to 20th year) is very profitable option even though the initial investment in considered high and pose high risk for crop failures (Moll, 1987). The net asset yields from oil plantation with the

Table 2 Land use pattern in PFR in 1995 and 2003 and its change use change pattern within the given period (a); and erosion rate for each given land use classes in 1995 and 2003 (b).

(a)

Land-use element	Area of land use class (ha)		Changes * (ha)
	1995	2003	
City	2130.627	2151.899	+ 21.272
Cemetery	87.302	90.126	+ 2.824
Quarry	119.468	119.586	+ 0.118
Mines	518.778	518.778	-
Sundry Non Tree Cultivation	3201.623	3207.957	+ 6.334
Sundry Tree Cultivation	34494.578	34633.937	+ 139.359
Grass	25087.623	25186.034	+ 98.411
Paddy	5832.417	5845.914	+ 13.497
Rubber	266873.894	267720.392	+ 846.498
Oil Palm	107734.269	107995.135	+ 260.866
Primary Forest	341094.066	339630.061	- 1464.006
Secondary Forest	22870.14	22944.966	+ 74.826
Total land area	810044.785	810044.785	

Note: *=changes of area in given classes reported as '+' for increment and '-' as reduction of area, respectively

(b)

Land use element	Erosion rate 1995 (ton/year)		Erosion rate 2003 (ton/year)	
	Erosion	mean	Erosion	mean
City	577.399	0.271	640.684	0.281
Cemetery	23.222	0.266	28.209	0.313
Quarry	28.672	0.24	29.059	0.243
Mines	129.695	0.25	113.326	0.257
Sundry Non Tree Cultivation	99535.257	31.089	117331.027	36.575
Sundry Tree Cultivation	31419455.36	910.852	37113380.55	1071.59
Grass	556493.65	22.182	657254.743	26.096
Paddy	1190921.227	204.19	1404328.845	240.224
Rubber	13800049.06	51.71	16286770.05	60.835
Oil Palm	7307507.732	67.829	8617903.778	79.799
Primary Forest	3502012.776	10.267	4102391.507	12.079
Secondary Forest	1348126.202	58.947	1591210.447	69.349
Total soil loss (ton/yr ⁻¹)	59224860.25		69891382.23	

Table 3 Erosion risk categories of land use classes in PFR.

Land-use class	Soil loss in 1995 group into risk categories [#]				Soil loss in 2003 group into risk categories [#]			
	1	2	3	4	1	2	3	4
City	2073.484	54.857	2.284		2133.71	18.189		
Cemetery	86.579	0.723			88.711	1.415		
Quarry	119.454	0.014			118.669	0.917		
Mines	516.624	2.154			495.511	23.267		
Sundry Non Tree Cultivation	3114.426	87.197			3120.242	87.715		
Sundry Tree Cultivation	34044.424	446.23	3.924		34169.557	463.384	0.996	
Grass	24926.969	159.121	1.533		24995.524	188.926	1.584	
Paddy	5809.636	22.781			5837.73	8.184		
Rubber	262174.189	4697.061	2.644		262786.969	4920.158	13.265	
Oil Palm	106045.166	1683.759	2.674	2.67	105905.887	2082.985	3.589	2.674
Primary Forest	339611.198	1478.97	3.9		338269.57	1355.207	5.284	
Secondary Forest	22661.477	208.664			22739.822	203.027	2.117	
Total area of erosion risk categories	800661.902	8841.531	14.675	2.67	800661.902	9353.374	26.835	2.674.

Note :

is the categories of erosion:

1 = Very low risk (0-1 ton ha⁻¹ yr⁻¹); 2 = Low risk (1-10 ton ha⁻¹ yr⁻¹);3 = Moderate risk (10-50 ton ha⁻¹ yr⁻¹) and 4 = Moderate high risk (50-100 ton ha⁻¹ yr⁻¹)

Table 4 Ecology service values and good in PFR for 1995

	Area (ha)	Ecological sevices							Ecological goods				Total rate (USD/ha)	Cost of individual landscape
		Water supply #	Erosion control #	Carbon stock *	Carbon sequestration *	Waste treatment #	Recreation #	Cultural and artistic info #	Food production #	Raw materials #	* Medical resources	Ornamental resources *		
Landscapes features 1995														
Tropical Primary Forest	341094.066	8	245	8	8		112	2	32	330	50	250	1045	356443298.97
Tropical secondary forest	22870.140	8	245	6	6		112		32	330		250	989	22618568.46
Grassland/ rangelands	25087.623						2		67	29			98	2458587.05
Rock (quarry and mines)	638.364									50			50	31918.20
Oil palm	107734.269	2								1200			1202	129496591.34
Rubber	266873.894	2								77			79	21083037.63
Paddy	5832.417								54				54	314950.52
Urban	2217.929												0	0.00
Other sundry crops cultivation	37696.201								54				54	2035594.85
Total land area	810044.903													
--Soil loss due erosion (ton/yr)	59224860.250												6	355349161.50
Total ecological services and goods in 1995													\$179,133,385.52	

ecological services values and goods based on Constanza *et al.* (1997)

* ecological services values and goods adopted from FAO (2003) – estimates of economic values of forest benefits.

Table 5 Ecology service values and good in PFR for 2003

	Area (ha)	Ecological services						Ecological goods				Total rate (USD/ha)	Cost of individual landscape	
		Water supply #	Erosion control #	Carbon stock *	Carbon sequestration *	Waste treatment #	Recreation #	Cultural and artistic info #	Food production #	Raw materials #	* Medical resources			Ornamental resources *
Landscapes features 2003														
Tropical Primary Forest	339630.061	8	245	8	8		112	2	32	330	50	250	1045	354913413.75
Tropical secondary forest	22944.785	8	245	6	6		112		32	330		250	989	22692392.37
Grassland/ rangelands	25186.034						2		67	29			98	2468231.33
Rock (quarry and mines)	638.364									50			50	31918.20
Oil palm	107995.135	2								1200			1202	129810152.27
Rubber	267720.392	2								77			79	21149910.97
Paddy	5845.914								54				54	315679.36
Urban	2242.025												0	0.00
Other sundry crops cultivation	37841.894								54				54	2043462.28
Total land area	810044.604													
--Soil loss due erosion (ton/yr)	69891382.230												6	419348293.38
Total ecological services and goods in 2003														\$114,076,867.13

ecological services values and goods based on Constanza et al (1997)

* ecological services values and goods adopted from FAO (2003) – estimates of economic values of forest benefits.

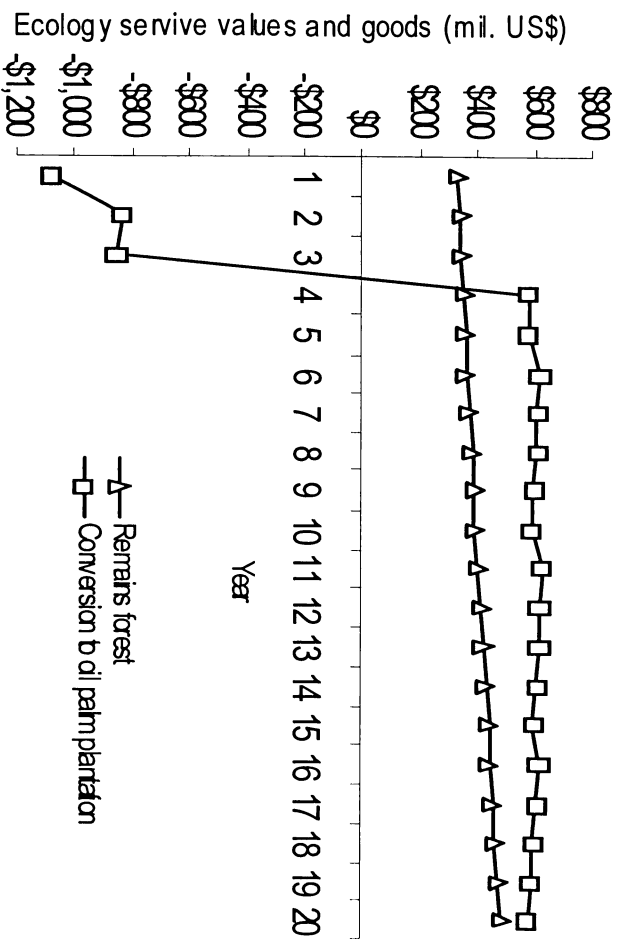
mean calculated from past 10 years market prices is US\$2374 ha⁻¹yr⁻¹, lucrative venture for improving if not strengthening rural economics. In the calculation of the production cost, a 2% compound interest for possibility of increment in cost from 1st to 20th year of plantation, and the risk of oil palm prices fluctuation in the global market. Subsequently, when full 17 parameters of ESGV are used, the total asset generated for ESGV indicated clearly that the ESGV assets are very much higher throughout the 20-year period. At the end 20th year, the asset generated from full ESGV worth sum of US\$ 963x10⁶ compared to US\$ 575x10⁶ generated from about 340x103 ha of oil palm plantation. At the same period, ESGV only valued at US\$481 millions when partial (11) parameters are considered. Adopting all the full ESGV parameters for sustainable landscape development in no doubt very important, however, the implementation is constrained by activities of agricultural land development activities for improving rural economics. Task of linking science-drive ESGV to the current development activities warrant detailed study of its own. In addition, science-drive ESGV making obligation of respective managers in development of landscape difficult to consider as the benefits is unforeseen. Empirical approach of calculating ESGV needs further refinement to its attributes taking into consideration local environmental and economic factors. This is to ensure ecological economics used is acceptable to stakeholders and sounds to decision makers as elaborated in Martin (2003). The gap between the ESGV and environmental policies in the Southeast Asian regions have in fact long been identified (Tomich *et al.*, 2004; Pattanayak, 2004) for degradation of ecological service values and goods in particular biodiversity and watershed functions apart from pollution. Mapping ESGV is one of the paramount efforts to assist in linking this gap, crucial for input into economical policy reviews towards environmental issues so as to ensure its participatory operational acceptance.

CONCLUSIONS

Study on the risk assessment due to landscape development using GIS has been undertaken in 3 terms (2002-2004) of Eco-Frontier Fellowship at NIES has been presented. The main aim the study is to assess risks due to the landscape development, which are divided into 5 aspects; (i) creation of spatial database from baseline information, (ii) structuring the simulation model based on spatial analysis in the GIS, and (iii) mapping ecological service values and goods, (iv) analyses of rate of landscape development from relation to generated impacts in term of changes in ESGV before and after development at PFR, and (v) analysis of simulated risks on development based on cost of ESGV. This report only last two tasks are highlighted, the others have been reported in the last two previous reports.

It is shown that changes that conversion of 1464ha (0.2% of PFR) of primary forest into agricultural activities and have significantly increased the erosion rate at total soil loss from 59 to 69mil ton/ha/yr. The mean soil loss for PFR is 0.8mil ton/ha/yr and if translated into ESGV term suppose every tonne cost US\$6, the soil loss costs about US\$4.8mil/yr. Majority of the resultant soil loss within all land use classes are within range of very low - low risk categories (less than 10 ton/ha/yr). ESV for PFR were costing US\$179x10⁶ in 1995, declined to US\$114x10⁶ in 2003 due to 0.2% reduction in forested area. The annual asset of converting 339,630ha primary forest into mass plantation determined from ESV cost less than original forest within period of 20yrs examined, i.e. at 20th year of conversion, the plantation and original forest cost US\$963 x10⁶ and US\$575 x10⁶, respectively. However, realistic parameters and compromising values of ecology services and goods used in PFR has yet to acceptable against stakeholders as well as sound to the policy makers as we anticipated such rapid assess of proposed landscape development only for modeling per se but its implementation is of practical for future sustainable landscape development.

(a)



(b)

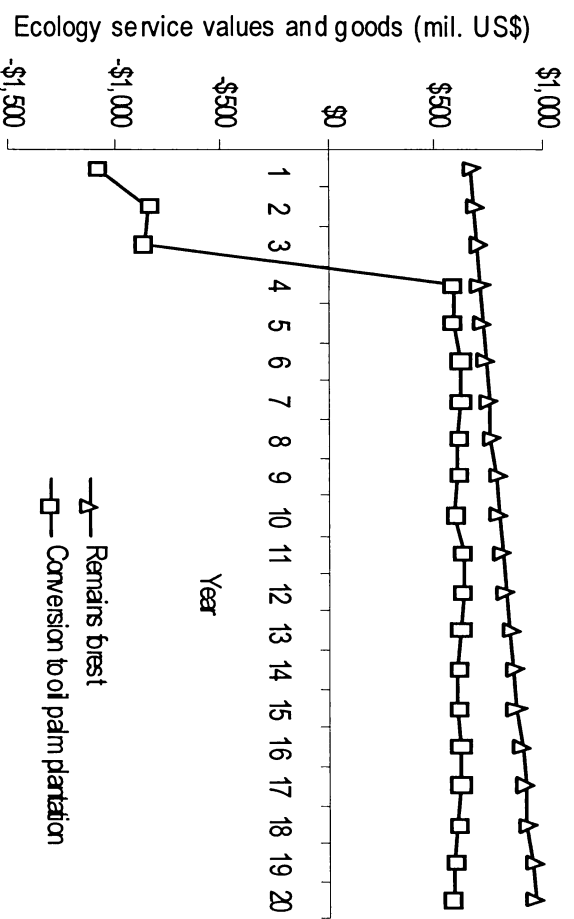


Fig. 2. Annual ecology service values and goods (ESVG) of keeping primary forest (339,630ha) intact versus converting it into mass oil palm plantation using: (a) 11 parameters of ecology service values and goods, and (b) the entire 17 recommended parameters of ecology service values and goods.

ACKNOWLEDGEMENTS

I deeply thank Dr Toshinori Okuda at NIES and the Environmental Agency's EFF Program for making my stay in Tsukuba possible. Thanks also due to all fellow scientists at Tropical Ecology Laboratory, NIES namely Dr. Shinya Numata, Mr. Sen Nishimura, Dr. Toshiaki Kondo and Ms. Mariko Suzuki, whom willingness to share their ideas and information to enable this study. AIRIES assistance and UTM's attachment release are also deeply appreciated.

REFERENCES

- Arif, S. Tengku Mohd Ariff, T. A.(2001). The case study on Malaysian oil palm. Procs of regional workshop on commodity export diversification and poverty reduction in South-east Asia. UNCTAD, Bangkok, Apr.3-5, 2001,15pp.
- Adger W. N., Brown, K., Cervigni, R., and Moran, D. 1995. Total economic value of forests in Mexico. *Ambio* 24: 286-296.
- ADB (Asian Development Bank) (2004). Regional and country highlight "South-east Asian Nations. URL: <http://www.adb.org/Countries/Highlights/SES.asp> (cited on Jan.18, 2005).
- Bishop, J. T. (ed) (1999) Valuing forest: A review of methods and application in developing countries. International Institute for Environment and Development, London.
- Baharuddin, K. (1988) Effects of logging on sediment yield in a hill dipterocarp forest in Peninsular Malaysia. *Journal of Tropical Forest Science*, 1(1):56-66.
- Bank Bumiputera-Commerce (2004), Special Industry Issue-Oil palm industry. Economic Research services, vol.4/2004. Bank Bumiputera-Commerce Sdn Bhd, Malaysia, 20pp.
- Bann, C., (1998) The economic valuation of tropical land use options : A manual for researchers.URL: <http://www.eepsea.org/publications/specialp2/ACF30F.html> (cited on Jan.18, 2005).
- Blamford, A., Bruner, A., Cooper, P., Constanza, R., Farber, S., Greem, R. E., Jenkins, M., Jefferies, P., Jessamy, V., Madden, J., Munro, K., Myers, N., Naeem, S., Paavola, J., Rayment, M., Rosendo, S., Roughgarden, J., Trumper, K., and Turner, R. T. (2002) Economic reasons for conserving wild nature. *Science*, vol 297, 950-953.
- Constanza, R., D'Arge, R., de Groot, R., Farber, S., Grasso, M., Hannon, B., Limburg, K., Naeem, S., O'Neil, R.V.,Paruelo, J., Raskin, R. G., Sutton, P. and Belt, M. V. D. (1997) The value of the world's ecosystem services and natural capital. *Nature*, vol 357, 253-260.
- de Groot, R. S., (1987) Environmental functions as a unifying concept for ecology and economics. *Environmentalist*,vol 7, 105-109
- de Groot, R. S., (1992) Functions of nature in environmental planning, management, and decision-making. Groningen: Wolters-Noordhoff.
- DTM (Dept of Treasury Malaysia) (2003). Economic Report 2002. Kuala Lumpur, Malaysia
- DSM (Dept of Statistics Malaysia) (2003). Malaysian Economic Statistics-Time series, Kuala Lumpur, Malaysia
- Douglas, I. (1968) Erosion in the Sg Gombak Catchment, Selangor, Malaysia. *Journal of Tropical Geography*, 26:1-26.
- FAO (Food and Agriculture Organization) (2003). Estimates of economic values of forest benefits. URL: <http://www.fao.org/DOCREP/003/W3641E18.htm> (cited on Jan.10, 2005).
- Fearnside,P.,(1997) Environmental services as a strategy for sustainable in rural Amazonia. *Ecol.Econ.*20(1), 53-70.
- Jalani,B.S.,Yusof,B., Ariffin,D., Chan, K.W. and Rajanaidu,N., (2002) Prospects of oil palm productivity: a Malaysian perspective, Malaysian Palm Oil Board, Kuala Lumpur, 9pp.

- Hashim, M. (2004) Risk Assessment on Landscape Development using Geographic Information System. Report of Eco-Frontier Fellowship 2004, AIRIES, Tokyo, Japan,
- Hashim, M. (2003) GIS Mapping of Ecological Services Value and Goods in Tropical Forest Ecosystem. Report of Eco-Frontier Fellowship 2003, AIRIES, Tokyo, Japan.
- Martin, J.S., (2003), Empiricism in ecological economics: a perspective from complex systems theory, *Ecological Economics* (46) no 3: 387-398
- MPOB (Malaysian Palm Oil Board) (2005). Statistics of annual average prices of oil palm products. URL: http://161.142.157.2/home2/home/ei_tajukmon02.html (cited on Jan.25, 2005).
- Mladenoff, D.J. and Baker, W.L., (1999), Development of forest and landscape modeling approaches. In Mladenoff, D.J. and Baker, W.L., (eds) *Spatial Modeling of forest landscape change : Approaches and applications*. Cambridge: Cambridge University Press, 1-14.
- Moll, H.A.J., (1987), *The economics of oil palm*. Pudoc Wageningen: Netherlands
- Morgan, R.P.C. (1974). Estimating regional variations in soil erosion hazards in Peninsular Malaysia. *Malayan Nature Journal* 28, 94-106
- NIES (2003). Japan-Malaysia Joint Research Project on Tropical Ecosystem http://www.nies.go.jp/biology/pasoh/English/current_issue.html
- Okuda, T., Suzuki, M., Adachi, N., Yoshida, K., Niiyama, K., Nur Supardi, M. N., Manokaran, N., and Hashim, M., (2003). Logging history and its impact on forest structure and species composition in the Pasoh Forest Reserve - Implication for the sustainable management of natural resources and landscapes-. Pp.15-34 in Okuda, T, Manokaran, N., Matsumoto, Y., Niiyama, K., Thomas, S. C., & Ashton, P. S. (eds.). "Pasoh - Ecology of a Lowland Rain Forest in Southeast Asia", Springer-Verlag, Tokyo.
- Okuda, T., Yoshida, K., Numata, S., Nishimura, S., and Hashim, M., (2003). Integrated ecosystem assessment - Towards sustainable natural resource use and management in Tropics. Pp.137-149 in Kobayashi, S., Matsumoto, Y., and Ueda, E., (eds.). *Rehabilitation of degraded tropical forests, Southeast Asia, Procs of the International Workshop on the landscape-level rehabilitation of degraded tropical forests.*, 18-19 Feb, 2003, Tsukuba, Japan.
- Pattanayak, S.K., (2004). Valuing watershed services: concepts and empirics from Southeast Asia. *Agriculture, Ecosystems & Environment*, 104(1):171-184.
- Peters, C. M., Gentry A. H., & Mendelsohn, R. O. 1989. Valuation of an Amazonian rainforest. *Nature* 339: 655-657.
- Pow, S.C., (2003), Perennial tree crop plantation systems in Malaysia: a model for productive efficient sustainable tree farms in developing countries in the Tropics. Procs of IFA-FAO Agriculture Conference-global food security and the role of sustainable fertilization, Rome, Italy, March 26-28, 2003.
- Pulzl, H. and Rametsteiner, E. (2002) Grounding international modes of governance into National Forest Programmes. *Forest Policy and Economics* 4(4):259-268
- Tan, G. and Shibasaki, R., (2003), Global estimation of crop productivity and the impacts of global warming by GIS and EPIC integration, *Ecological Modelling*, (15): 357-370.
- Tomich, T.P., K.Chomitz, H. Francisco, A.N. Izac, D. Murdiyarso, B.D. Ratner, D. E. Thomas and M. van Noordwijk. (2004), *Environmental Services and Land Use Change: Bridging the Gap between Policy and Research in Southeast Asia*. *Agriculture, Ecosystems & Environment*, 104(1):5-18.
- Torras, Mariano (2000) The total economic value of Amazonian deforestation 1978-1993. *Ecological economics* 33, 283-297.
- Turner, R.K., (1988) *Economics, growth and sustainable environments*. In Collard, D. (ed). London: Macmillan.
- Zulkifli, Y (2005) Report of EFF-2005, In press.